72 particle analyzer was used to monitor changes in particle sizes. A conventional ESP placed at the

73 outlet of the agglomeration chamber was used to collect particles. This experiment investigated

74 the effect of using a coupling field compared with a single field, in terms of median particle

75 diameter change and removal efficiency.

76

EXPERIMENTAL SETUP AND METHODOLOGY

78

79 In this experiment, the feeding system comprised a SAG 410/L aerosol generator, with a flow

80 rate of 1–264 g h^{-1} (Fig. 1). The agglomeration chamber was a vertical cuboid fabricated from

81 polymethyl methacrylate (PMMA). The height of the chamber was 1200 mm, and the side length

82 was 120 mm. Inside the chamber, four stainless steel wire electrodes were located every 100 mm

83 and each had a diameter of 2 mm and a length of 80 mm. Additionally, the chamber contained

84 two discharge plates for the production of pulse corona discharges. Sound-absorbing foam was

85 used in the chamber to reduce the reflection of sound waves. The acoustic wave system

86 comprised a SFG-1013 signal generator, YF-513 compression driver, and a QSC RMX2450

87 power amplifier, which were used to set acoustic frequencies between 180 and 5500 Hz. Based

88 on the consideration that energy input breaks aggregates into pieces, SPLs between 135 and 143

89 dB were chosen in this study. These SPLs have been proven effective in other studies (Liu *et al.*,

90 2009; Wang *et al.*, 2011). The pulsed corona discharge system involved a narrow-pulse HV

91 generator, Tektronix TCP0150 current probe, Tektronix DPO4034 digital oscilloscope, and

92 P150-GL HV probe. The system was used to control input parameters and acquire the I-V curve.

93 Subsequently, particles flowed through the agglomeration chamber, and particle size distribution

94 (PSD$_1$) was captured using a DP-02 laser particle size analyzer. An electrical low pressure

95 impactor (ELPI) was placed after the ESP to measure the particle concentration and particle size

96 distribution between 0 and 10 μm (PSD$_2$). A two-stage diluter was used to ensure that the flow

97 rate to the ELPI was 10 m^3 h^{-1}. The ESP was wire-plated and used a negative direct current (DC)

98 power supply.

99 Jiaxing coal-fired plant fly ash particles were used in this experiment, and they were mixed

100 before being released into the agglomeration chamber. A DP-02 laser particle size analyzer was

101 used to ensure that a laser beam could penetrate the bottom part of the agglomeration chamber. In

102 this study, the ELPI measurement range was less than10 μm, but a substantial proportion of the

103 agglomerated particles was larger than 10 μm. The DP-02 laser particle size analyzer was utilized

104 to measure the size change of the agglomerated particles to verify the effect of the coupling field.

105

106 **THEORETICAL ANALYSIS**

107

108 Particle charges, particle collisions, and particle–gas interactions are involved in the processes

109 of particle agglomeration and removal. Thus, when particles flow through an agglomeration

110 chamber, multiple forces may act upon them, such as Coulomb force between charged particles,

111 viscous Stokes' drag force, gravitational force, inertial force, and buoyancy force. The mechanism

112 resulting from these mutually influential forces is complex.

113

114 *Pulse Corona Discharge*

115   High-voltage pulse discharge differed from negative DC discharge; the high-voltage discharge

116   process comprised two stages: a discharge period and an intercritical period. In this study,

117   high-voltage generator frequency ranged from 100 to 300 Hz, but the pulse width was only

118   300–400 ns, and the intercritical period was much longer than the discharge period. Discharge

119   occurred in formed streamer-channels in which numerous high-energy free electrons and ions

120   coexisted. The high-energy free electrons migrated more rapidly than did the ions, and in the

121   agglomeration chamber, the concentrations of positive and negative ions were extremely

122   unbalanced (McAdams, 2007). Thus, the discharge period, mainly involved the charging of fine

123   particles by high-energy free electrons. After the discharge period, numerous ions remained in the

124   reaction space. The number of negative ions was substantially lower than that of positive ions,

125   and most of the negative ions were neutralized within a short period (McAdams, 2007). Thus,

126   charges from positive ions were more prevalent on fine particles.

127   Based on the preceding analysis, during the discharge period, electron flux was solved to

128   identify charge number (Liu and Kapadia, 1978; O'Hara *et al*., 1989):

129

130    $\Pi|_r = -\mu_e N_e(r)E - D_e \nabla N_e(r)$ (1)

131

132   where μ_e is the migration rate of high-energy free electrons; $N_e(r)$ is the concentration of

133   high-energy free electrons r away from the particle center; D_e is the diffusion coefficient of

134   high-energy free electrons; and $\nabla N_e(r)$ is the concentration gradient of high-energy free

135 electrons r away from the particle center.

136 When electron flux is small, Einstein's equations should be introduced to solve Eq. (1), where

137 the energy distribution of high-energy free electrons is assumed to conform to Maxwell

138 distribution (Bates *et al.*, 1962):

140 $$KT_e = \frac{2}{3} X(\frac{D_e}{\mu_e})e \qquad (2)$$

142 where K is the Boltzmann constant; $K = 1.38 \times 10^{-23}$ J K^{-1}; T_e is the temperature of high-energy

143 free electrons; X is a constant; and $X = 3/2$ when energy distribution of high-energy free electrons

144 conforms to Maxwell distribution,

146 $$\ln N_{e(r)} = \frac{e\Phi_{(r)}}{KT_e} + A \qquad (3)$$

148 Where $\Phi_{(r)}$ is electric potential r away from the particle center; A is a constant; and the

149 boundary conditions are $\Phi_{(r)} = 0$, $N_{e(r)} = N_{e0}$,

151 $$N_{e(r)} = N_{e0} \exp(\frac{-e\Phi_{(r)}}{KT_e}) \qquad (4)$$

153 where N_{e0} is the original concentration of high-energy free electrons. If electron flux is large

154 enough or the ratio of electric field intensity to gas molecular density is large enough, the energy

155 distribution of high-energy free electrons deviates from Maxwell distribution; Eq. (4) is no longer

156 applicable; and $\Phi_{(r)}$ refers to electric potential under the influence of particles' own charge and

157 external electric field,

158

159 $$\Phi_{(r)} = \frac{-q_1}{4\pi\varepsilon_0 r} - [\frac{r}{R} - (\frac{\varepsilon - \varepsilon_0}{\varepsilon + 2\varepsilon_0})\frac{R^2}{r^2}]E_0 a\cos\theta \qquad (r \geq R) \tag{5}$$

160

161 where q_1 is particle charge; r is the distance between a point in the reaction system and a

162 particle center; ε is a particle's relative dielectric constant; ε_0 is the permittivity of vacuum; E_0 is

163 the original electric field intensity in the reaction system; θ is the angle of a line between the

164 particle center and a point r away from the particle center and the electric field line.

165 Eqs. (1, 4, 5) indicate that high-energy free electron flux could be created on the surface of

166 particles,

167

168 $$\frac{dq_1}{dt} = -2\pi ea^2 \int_{\theta=0}^{\theta=\pi} \Pi|_{r=a} \sin\theta d\theta \tag{6}$$

169

170 To solve Eq. (6), the boundary condition ($\nabla N_e(r=R)=0$) should be applied, because electron

171 density distribution conforms to Boltzmann distribution, and continuous electron distribution

172 occurs around particles. Thus, solution is as follows:

173

174 $$q_1 = \frac{4\pi\varepsilon_0 RKT_e}{e}\ln[1 + \frac{N_{e0}Re^2\mu_e E_0 t}{2\varepsilon_0 KT_e}(\frac{e^\alpha - e^{-\alpha}}{\alpha})] \tag{7}$$

175

176
$$\alpha = \frac{-3eE_0 R}{(\frac{\varepsilon}{\varepsilon_0} + 2)KT_e}$$
(8)

177

178   With the increase of electric field intensity, particles receive more charges. Similarly, after the

179   discharge period, particles capture positive ions and become positively charged:

180

181
$$q_2 = \frac{4\pi\varepsilon_0 RKT}{e} \cdot \ln[1 + \exp(-\frac{e^2 B}{RKT})\frac{\pi u_i e^2 RN_{i0}t}{KT}]$$
(9)

182

183   where B is a constant related to R, ε, and ε_0.

184   Due to the opposite charges present among particles during the discharge period, Coulomb

185   force exists in the agglomeration chamber:

186

187
$$F_C = \frac{q_1 q_2}{4\pi\varepsilon R^2}$$
(10)

188

189   According to studies on fine particle charges (Jiang *et al.*, 2015; Xu *et al.,* 2009), small and

190   large particles exhibit opposite charges and attract each other, resulting in the formation of

191   numerous aggregates.

192

193   ***Acoustic Wave Field***

194   Acoustic wave is a promising pretreatment method for achieving particle agglomeration. In an

195   intensive acoustic wave field, particles are entrained by gas media. Orthokinetic interaction plays

196   a key role in the acoustic agglomeration process (Sheng and Shen, 2007). In the orthokinetic

197   interaction process, an acoustic wave causes air medium vibration; because of the viscosity of the

198   gas, particles in the aerosol vibrate in response. Large particles with large inertia are not readily

199   carried by gas while small particles are easy to be carried. Different vibration amplitudes result in

200   relative motion and collision between particles of different sizes, and the entrainment coefficient

201   μ_p can be defined as follows (Zhou $et\ al.$, 2017):

202

203    $$\mu_p = \frac{u_p}{u_g}$$ (11)

204

205   Several forces are present in the acoustic field, namely pressure gradient force, viscous Stokes'

206   drag force, and the force required for fluid acceleration near the particles. In this study, particle

207   motion was limited within Stokes flow (Re < 1), and only viscous Stokes' drag force was

208   considered:

209

210    $$F_p = \frac{6\pi\mu_g R(u_g - u_p)}{c_c}$$ (12)

211

212   where μ_g is the dynamic viscosity of gas; R is particle radius; u_g is the velocity of gas; u_p is the

213   velocity of particles; and C_c is Cunningham correction factor (Fan $et\ al.$, 2013).

214   Entrained particles in gas medium vibrate and lag behind the gas. The particle motion equation

215   can be expressed as follows:

216

217 $$u_g = u_0 \sin \omega t \qquad (13)$$

218

219 The velocity of particles can then be simplified as follows,

220

221 $$u_p = \frac{u_0 \sin(\omega t - \varphi)}{\sqrt{1 + \omega^2 \tau^2}} \qquad (14)$$

222

223 where τ is the relaxation time of particles, u_0 is the velocity amplitude of gas medium; ω is

224 angular frequency; and $\omega = 2\pi f$,

225

226 $$\tau = \frac{2 \rho R^2 C_c}{9 \mu_g} \qquad (15)$$

227

228 Where f is the acoustic wave frequency. φ is the phase difference between gas medium and

229 particles, so the entrainment coefficient μ_p can be de represented as follows:

230

231 $$\mu_p = \frac{1}{\sqrt{1 + \omega^2 \tau^2}} = \frac{1}{\sqrt{1 + 4\pi^2 f^2 \tau^2}} = \cos \varphi \qquad (16)$$

232

233 Fig. 2 presents the effects of particle size and acoustic wave frequency on the acoustic

234 entrainment factor. The value of μ_p is between 0 and 1, and a larger entrainment factor signifies

235 that particles are readily entrained by the gas medium. When μ_p reaches 1, the vibration

236 amplitude of the particles and air medium is synchronized. By contrast, when μ_p is 0, particle

237 motion is not affected by acoustic waves and remains vibrationless. Larger particles are easily

238 entrained when acoustic wave frequency is low. Specific frequencies, correspond with specific

239 entrainment factors. These factors can be utilized to prove the existence of optimal acoustic

240 parameters for inducing relative movement and collision of particles with various diameters.

241 In the acoustic wave field, countless particles collide, and large particles are likely to be

242 shattered by acoustic waves. Because of the existence of a pulse corona field, particles with

243 various diameters receive opposite charges due to differences in charging mechanisms, and

244 Coulomb force may cause them to adhere and aggregate into larger particles. Thus, the coupling

245 effect increases the possibility of reunion and agglomeration. Application of ESP substantially

246 increases the efficiency of small particle removal.

247

248 **RESULTS AND DISCUSSION**

249

250 In this study, fly ash particle was assumed to be a convex spherical particle. A DP-02 laser

251 particle analyzer and ELPI were utilized to measure the particle agglomeration process and

252 particle removal efficiency respectively. Studies have focused on measuring removal efficiency

253 using ELPI. ELPI measurement range is less than 10 μm; thus, ELPI is insufficient for

254 monitoring particles with larger diameters. In this study, DP-02 was selected to measure particle

255 size change online. The particle size measuring mechanisms of ELPI and the DP-02 laser particle

256 analyzer differ; ELPI data are based on aerodynamic size, whereas DP-02 data are based on

257   optical particle size. As presented in Fig. 3, original particle size distribution PSD_1 (DP-02) was

258   compared with transformed PSD_2 (ELPI) in the same coordinate system. The results show that the

259   curves of the PSDs coincided; thus, the respective PSDs could be used to investigate

260   agglomeration and removal efficiency.

261   In a DC electric field, I–V characteristic curve was used to analyze the power supply properties.

262   A narrow-pulse HV generator was used as an intermittent discharge device. Consideration of

263   energy input was required. Energy input reflects electron and ion density changes and reaction

264   processes, as well as migration rate. Single pulse energy was introduced to describe pulsed

265   corona discharge energy consumption. Pulse corona discharge frequency ranged from 100 to 300

266   Hz. Fig. 4 illustrates the pulsed corona discharge characteristics. When input voltage increased,

267   single pulse energy will increased; in other words, more energy was injected into the reaction

268   space, and 100 Hz single pulse energy was the largest, followed by 200 and 300 Hz.

269

270   ***Effects of Pulse Corona Discharge on the Agglomeration of Fine Particles***

271   The purpose of this experiment was to verify the influence of pulse corona discharge on PSD_1.

272   The results are as presented in Fig. 5(a). New peak values in the pulse field were noted, and the

273   corresponding median particle size increased. When the voltage was 35 and 45 kV, the new peak

274   value after agglomeration was approximately 30 μm, and the volume proportions of 27-μm

275   particles reached 4.46% and 11.84%, respectively. Most small particles remained. When the pulse

276   voltage increased to 55 kV, the volume proportion of small particles dramatically decreased. The

277   proportion of 9.82-μm particles dropped to 31.28%, and the proportion of 27-μm particles

278   increased to 25.96%. When the pulse frequency was 100 or 200 Hz, new peaks appeared at 155

279   μm, and the peak values were 6.29% and 4.01%, respectively. Fig. 5(b) illustrates that in the

280   pulse corona discharge field, the median particle size after agglomeration clearly increased. As

281   input voltage increased and pulse frequency decreased, D_{50} increased, reaching 23 μm under the

282   condition of 55 kV-100 Hz. Energy input ionized the air and the concentration of high-energy free

283   electrons and ions increased dramatically, which may have increased the intensity of mutual force

284   between charged particles, resulting in easier aggregation of particles.

285   Agglomeration ratio (R_{ae}) was defined as the ratio of the median particle size after

286   agglomeration to the median size of the original particles. As presented in Fig. 6, when the input

287   voltage was 55 kV, the pulse frequency was 100 Hz; the corresponding median diameter was

288   23.39 μm; and R_{ae} was 2.93. When the input voltage was increased from 35 to 45 kV, D_{50} and R_{ae}

289   did not exhibit significant changes. This finding suggested that large pulse voltage and low pulse

290   frequency should be selected to maximize the effectiveness of agglomeration within the limits of

291   the experimental conditions.

292

Effects of Acoustic Wave on the Agglomeration of Fine Particles

294   In these experiments, because of limits in the experimental conditions, SPLs were 135, 139,

295   and 143 dB. As denoted in Table 1, when SPL was 143dB, both the D_{50} and R_{ae} both first

296   increased and then decreased as the acoustic wave frequency increased. With the decrease of SPL,

297   D_{50} and R_{ae} also decreased. Therefore, the larger the SPL is, the more intense the particle

298   vibration. Additionally, the probability of collision between particles increased and the optimum

299 acoustic frequency was 1600 Hz.

300 Experiments were performed to investigate the effects of acoustic waves on the agglomeration

301 of fine particles. According to the results of the previously described experiments, SPL = 143 dB

302 was selected. As Fig. 7 presents, the effect of the acoustic frequency on the aggregation of fine

303 particles was complex. When acoustic frequency was 2400 Hz, a new small peak value appeard

304 between 20 and 30 μm. The effects of 800 and 1200 Hz were similar: a wide range of peaks

305 appeared; almost all fine particles agglomerated into large particles; and most particles were 20 –

306 80 μm. The frequency of 1600 Hz was most effective for agglomeration; not only did the small

307 particles all agglomerate into large particles but also the peak value was larger than 40 μm after

308 agglomeration. Additionally, obvious peaks were evident at 200 μm. when SPL was 143 dB, the

309 acoustic frequency was 1600 Hz; D_{50} of fine particles changed from 7.98 to 57.56 μm, and R_{ae}

310 reached 7.21 (Fig. 8). By contrast, when acoustic frequency was 2400 Hz, D_{50} was only 20.52 μm,

311 and the corresponding R_{ae} was only 2.57. The findings suggested that acoustic frequency should

312 neither be too large nor too small and that optimal frequency value was 1600Hz. Orthokinetic

313 interaction was the predominant mechanism, suggesting that relative movement and collision

314 resulted from differences in the entrainment degrees of particles. If the acoustic frequency was

315 too small, all particles vibrated with the gas medium, and relative motion among particles with

316 different sizes was small. If the acoustic wave frequency was too large, all particles remained still

317 and could not vibrate with the gas medium, and a relative motion effect was not obvious.

318 Therefore, selection of an appropriate acoustic frequency was crucial for ensuring an optimal

319 fine particle agglomeration effect.

320

Effect of Coupling Field on the Agglomeration of Fine Particles

This experiment was conducted in the acoustic wave and pulse discharge coupling field. The results are presented in Fig. 9. The pulse frequency and voltage were 100 Hz and 55 Kv, respectively, and the acoustic frequency and SPL were 1600 Hz and 143 dB, respectively. The volume proportion of large particles obviously increased, whereas that of fine particles decreased substantially. Most particles with diameters of 9.82 μm were agglomerated into larger particles. The volume proportion of particles with a diameter of 200.68 μm increased to 16.98% in the coupling field and to 1.93% in the pulse corona discharge field. As denoted in Fig. 9(b), D_{50} and R_{ae} significantly increased in the coupling field compared with the single field, reaching 73.28 μm and 9.18, respectively. Thus, the efficiency of fine particle agglomeration in the coupling field was greater than that of the single field, and D_{50} increased considerably.

Particle Removal Efficiency in Coupling Field

To determine whether agglomeration in the coupling fieldincreased removal efficiency, a laboratory-scale ESP with a negative DC supply was placed at an outlet of the agglomeration chamber. Studies have proven that a negative DC can remove particles efficiently, with the exception of fine particles. In this study, acoustic wave and pulse corona discharge coupling field were used as a pretreatment to force fine particles to agglomerate into larger particles, resulting in significantly more efficient fine particle removal. Removal efficiency was calculated using Eq. (17):

341

342 $$\eta = 1 - \frac{N}{N_0}$$ (17)

343

344   where N_0 is the initial particle number concentration (PNC), and N is PNC at the outlet of the

345 ESP. In this study, the parameters were the same as those used in the previously described

346 agglomeration experiment. DC supply voltage was regulated to 16 kV, which was the highest

347 voltage applied to the ESP in this experiment. As shown in Fig. 10, PNC clearly decreased after

348 particles passed through the ESP. Additionally, the amplitude of PNC decreased less for particles

349 with diameters of 0.15 μm than for those of other sizes. As Fig. 11 illustrates, ESP removal

350 efficiency significantly increased with pretreatment, especially in the coupling field. When only

351 DC supply was applied, removal efficiency was 80.7%, and that of particles with diameters of

352 0.15 μm was less than 75%.

353   When particles passed through the agglomeration chamber first, fine particles removal

354 efficiency increased sharply. When the pretreatment was pulse corona discharge, the efficiency

355 increased to approximately 91.6%, and when an acoustic wave field was applied, removal

356 efficiency further increased to 94.4%. Particles with diameters of 0.15 μm were present in the

357 critical area of field charge and diffusion charge; although these particles were difficult to remove,

358 pretreatment resulted in considerable increases in the efficiency of their removal from 75% to

359 80%, 85%, and 91%. Additionally, overall particle removal efficiency reached 98.3%.

360

361 **CONCLUSIONS**

362

363   Comparative experimental studies were conducted to verify the effects of pulse corona

364 discharge and acoustic wave coupling on particle agglomeration and removal efficiency. The

365 coupling field clearly increased agglomeration efficiency when the pulse corona discharge and

366 acoustic wave parameters were 55 kV-100 Hz and 143 dB-1600 Hz, respectively, and D_{50} and R_{ae}

367 values were 11.23 μm, 1.51 and 57.56 μm, 7.21. In the coupling field, the corresponding values

368 increased to 73.28 μm and 9.18, and the effect on agglomeration was obvious. The acoustic wave

369 value of 1600 Hz was optimal for enhancing the agglomeration effect, and when the SPL value

370 rose, agglomeration efficiency increased in response. When pulse frequency was 100 Hz, the

371 single pulse energy injection reached its maximum, and more fine particles were charged and

372 aggregated into larger particles. Under the acoustic wave and pulse corona discharge coupling

373 field, overall removal efficiency reached 98.3%, and particles with diameters of 0.15 μm were in

374 the critical area of field charge and diffusion charge. Although these particles were difficult to

375 remove, the grade removal efficiency substantially improved, from 75.1% to 80%, 85.3%, and

376 91.4%. Finally, the coupling field was more effective than the single field as a pretreatment for

377 fine particle.

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REFERENCES

385

386 Bates, D.R., Massey, H.S.W. and Amdur, I. (1962). Atomic and molecular processes. *Phys.*

387   *Today.* 15: 63-63.

388 Chen, C.C., Tao, C.J. and Cheng, H.C. (2002). Condensation of supersaturated water vapor on

389   charged/neutral nanoparticles of glucose and monosodium glutamate. *J. Colloid Interface Sci.*

390   255: 158-170.

391 Chen, H., Wang, T., Luo, Z., Zhou, D., Lu, M., He, M., Fang, M. and Cen, K. (2017).

392   Agglomeration kernel of bipolar charged particles in the presence of external acoustic and

393   electric fields. *Aerosol Air Qual. Res*. 17: 857-866.

394 Fan, F., Yang, X. and Chang, N.K. (2013). Direct simulation of inhalable particle motion and

395   collision in a standing wave field. *J. Mech. Sci. Technol.* 27: 1707-1712.

396 Hoffmann, T.L. (2000). Environmental implications of acoustic aerosol agglomeration.

397   *Ultrasonics.* 38: 353.

398 Jiang, J.P., Luo, Z.Y., Chen, H., Zhou, D., Sha, D.H., Fang, M.X. and Cen, K.F. (2015).

399   Orthogonal design process optimization for particle charge distribution of mosquito coil smoke

400   aerosol enhanced by pulsed corona discharge. *Powder Technol.* 286: 507-515.

401 Le, T.C., Lin, G.Y. and Tsai, C.J. (2013). The Predictive Method for the Submicron and

402   Nano-Sized Particle Collection Efficiency of Multipoint-to-Plane Electrostatic Precipitators.

403   *Aerosol Air Qual. Res*. 13: 1404–1410.

404   Liao, Z., Li, Y., Xiao, X., Wang, C., Cao, S. and Yang, Y. (2018). Electrostatic precipitation of

405   submicron particles with an enhanced unipolar pre-charger. *Aerosol Air Qual. Res*. 18.

406   Liu, B.Y.H. and Kapadia, A. (1978). Combined field and diffusion charging of aerosol particles

407   in the continuum regime. *J. Aerosol Sci*. 9: 227-242.

408   Liu, J., Wang, J., Zhang, G., Zhou, J. and Cen, K. (2011). Frequency comparative study of

409   coal-fired fly ash acoustic agglomeration. *Res. J. Environ. Sci*. 23: 1845-1851.

410   Liu, J., Zhang, G., Zhou, J., Wang, J., Zhao, W. and Cen, K. (2009). Experimental study of

411   acoustic agglomeration of coal-fired fly ash particles at low frequencies. *Powder Technol*. 193:

412   20-25.

413   Liu, J.X., Gao, J.H., Gao, J.M., Wang, X.F. and Wu, S.H. (2009). Agglomeration of particles

414   during coal combustion in multistage spouted fluidized tower. *Korean J. Chem. Eng*. 26:

415   907-912.

416   Luo, Z., Chen, H., Wang, T., Zhou, D., Lu, M., He, M., Fang, M. and Cen, K. (2017).

417   Agglomeration and capture of fine particles in the coupling effect of pulsed corona discharge

418   and acoustic wave enhanced by spray droplets. *Powder Technol*. 312: 21-28.

419   Matsoukas, T. (1997). The coagulation rate of charged aerosols in ionized gases. *J. Colloid

420   Interface Sci*. 187: 474-483.

421   McAdams, R. (2007). Pulsed corona treatment of gases: System scaling and efficiency. *Plasma

422   Sources Sci. Technol*. 16: 703.

423   O'Hara, D.B., Clements, J.S., Finney, W.C. and Davis, R.H. (1989). Aerosol particle charging by

424   free electrons. J. *Aerosol Sci*. 20: 313-330.

425 Sheng, C. and Shen, X. (2007). Simulation of acoustic agglomeration processes of poly-disperse

426 solid particles. *Aerosol Sci. Technol.* 41: 1-13.

427 Wang, J., Liu, J., Zhang, G., Zhou, J. and Cen, K. (2011). Orthogonal design process

428 optimization and single factor analysis for bimodal acoustic agglomeration. *Powder Technol.*

429 210: 315-322.

430 Wang, Y., Cheng, K., Tian, H.Z., Yi, P. and Xue, Z.G. (2016). Emission characteristics and

431 control prospects of primary pm2.5 from fossil fuel power plants in china. *Aerosol Air Qual.*

432 *Res.* 16: 3290-3301.

433 Xu, F., Luo, Z., Bo, W., Zhao, L., Gao, X., Fang, M. and Cen, K. (2009). Experimental

434 investigation on charging characteristics and penetration efficiency of PM2.5 emitted from coal

435 combustion enhanced by positive corona pulsed esp. J. *Electrostat.* 67: 799-806.

436 Xu, P., Wang, W., Ji, J. and Yao, S. (2014). Analysis of the contribution of the road traffic

437 industry to the PM2.5 emission for different land-use types. *Comput Intell Neurosci.* 2014:

438 821973.

439 Yan, J., Chen, L. and Yang, L. (2016). Combined effect of acoustic agglomeration and vapor

440 condensation on fine particles removal. *Chem. Eng. J.* 290: 319-327.

441 Zhang, G., Zhang, L., Wang, J. and Hu, E. (2017). Improving acoustic agglomeration efficiency

442 by addition of sprayed liquid droplets. *Powder Technol.* 317: 181-188.

443 Zhang, G.X., Liu, J.Z., Wang, J., Zhou, J.H. and Cen, K.F. (2012). Numerical simulation of

444 acoustic wake effect in acoustic agglomeration under Oseen flow condition. *Chin. Sci. Bull.* 57:

445 2404-2412.

446 Zhou, C.H., Gong, S., Zhang, X.Y., Liu, H.L., Xue, M., Cao, G.L., An, X.Q., Che, H.Z., Zhang,

447 Y.M. and Niu, T. (2012). Towards the improvements of simulating the chemical and optical

448 properties of Chinese aerosols using an online coupled model – CUACE/Aero. *Tellus B Chem*

449 *Phys Meteorol*. 64: 91-102.

450 Zhou, D., Luo, Z., Fang, M., Lu, M., Jiang, J., Chen, H. and He, M. (2016a). Numerical

451 calculation of particle movement in sound wave fields and experimental verification through

452 high-speed photography. *Appl. Energy*. 185: 2245-2250.

453 Zhou, D., Luo, Z., Jiang, J., Chen, H., Lu, M. and Fang, M. (2016b). Experimental study on

454 improving the efficiency of dust removers by using acoustic agglomeration as pretreatment.

455 *Powder Technol*. 289: 52-59.

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467 **Table 1.** Median particle diameters and agglomeration ratios of particles under various acoustic

468 SPLs and frequencies.

frequency	parameters	SPL		
		135dB	139dB	143dB
800Hz	$D_{50}/\mu m$	14.69	24.99	36.88
	R_{ae}	1.84	3.13	4.62
1200Hz	$D_{50}/\mu m$	18.76	27.22	36.95
	R_{ae}	2.35	3.41	4.63
1600Hz	$D_{50}/\mu m$	28.50	41.27	57.56
	R_{ae}	3.57	5.17	7.21
2000Hz	$D_{50}/\mu m$	23.15	26.10	28.10
	R_{ae}	2.90	3.27	3.52
2400Hz	$D_{50}/\mu m$	9.50	15.97	20.52
	R_{ae}	1.19	2.00	2.57

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Figure Captions

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484   **Fig. 1.** Schematic of experimental setup.
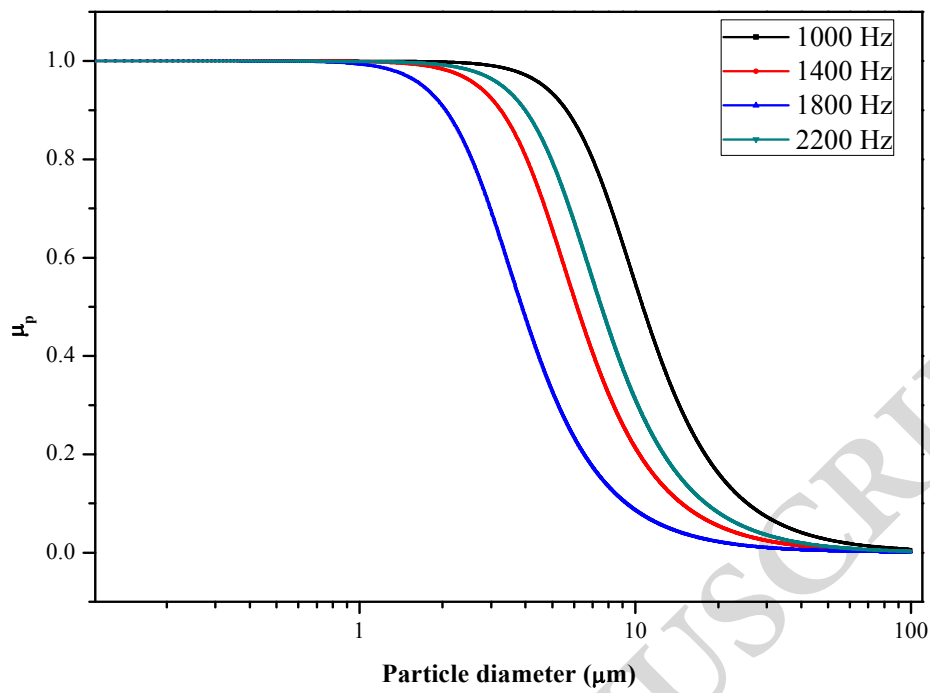
485   **Fig. 2.**Relationship between the entrainment factor and particle diameter at various acoustic

486   frequencies.

487   **Fig. 3.**Original particle size distribution of the ash.

488   **Fig. 4.** Influence of the input voltage on the single pulse energy of pulse corona discharge.
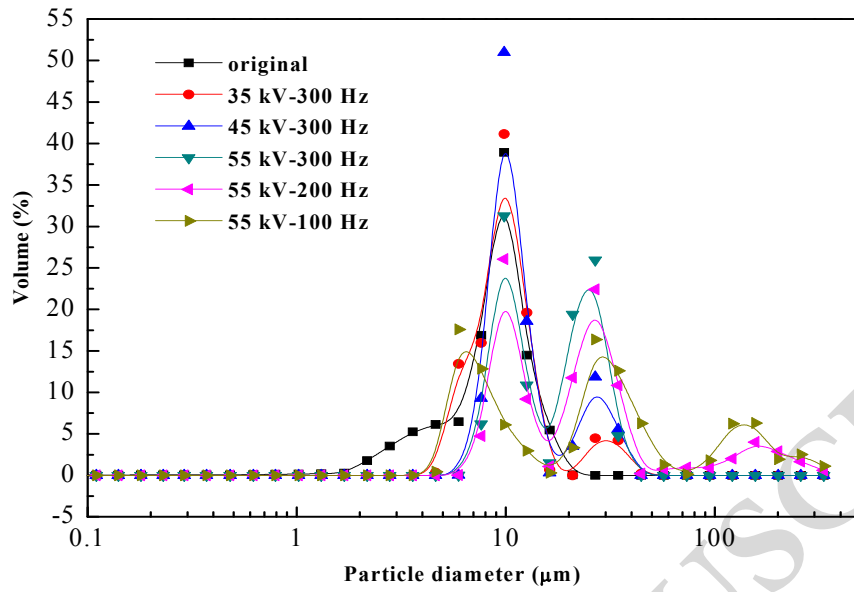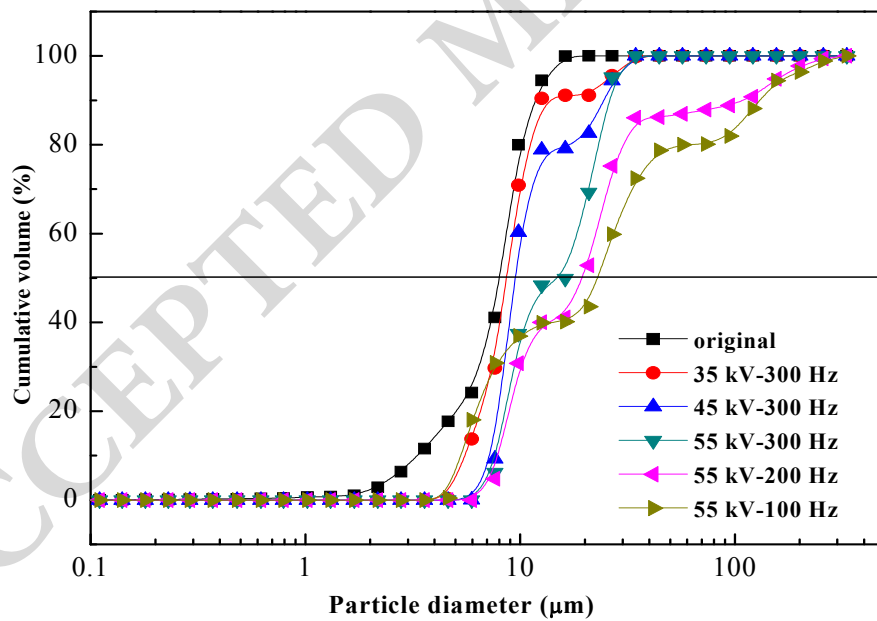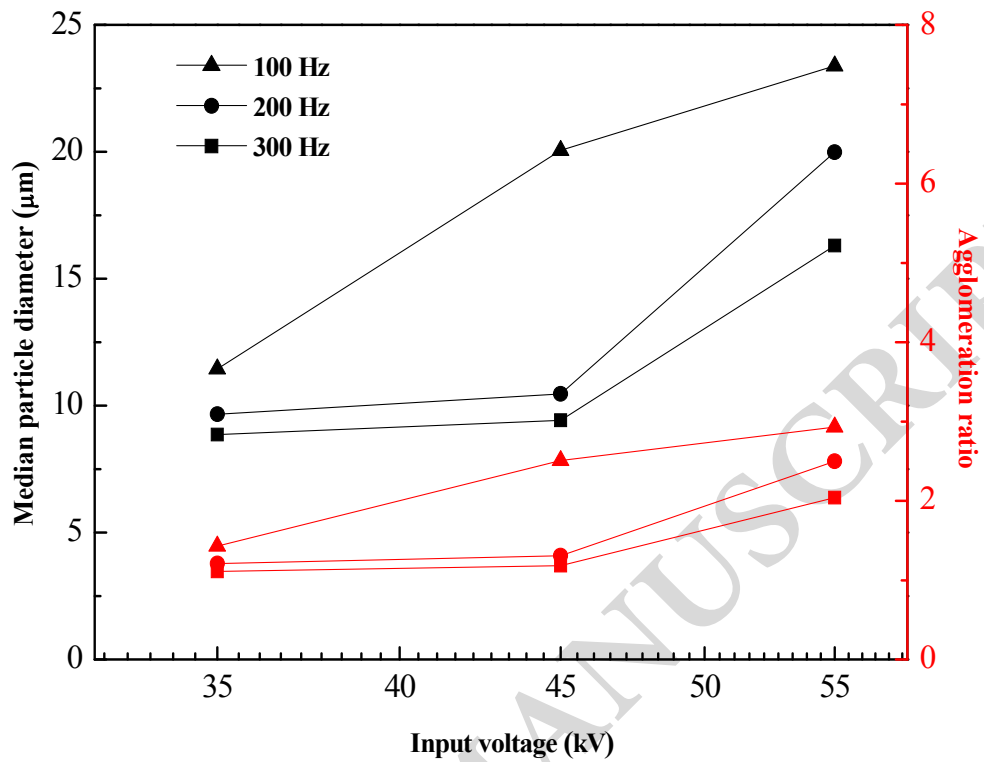
489   **Fig. 5.** Volume differential (a) and cumulative distribution (b) of particle size under various pulse

490   input voltages and frequencies.

491   **Fig. 6.** Median particle diameters and agglomeration ratios of particles under various pulse input

492   voltages and frequencies.

493   **Fig. 7.** Volume differential (a) and cumulative distribution (b) of particle size under various

494   acoustic wave frequencies (SPL = 143 dB).

495   **Fig. 8.** Median particle diameters and agglomeration ratios of particles under various acoustic

496   wave frequencies (SPL = 143 dB).

497   **Fig. 9.** Volume differential (a) and cumulative distribution (b) of particle size under single field

498   and coupling field.

499   **Fig. 10.** Comparison of particle number distribution after ESP under various pretreatment

500   conditions.

501   **Fig. 11.** Grade removal efficiency of fine particles after application of ESP with single field and

502   coupling field.

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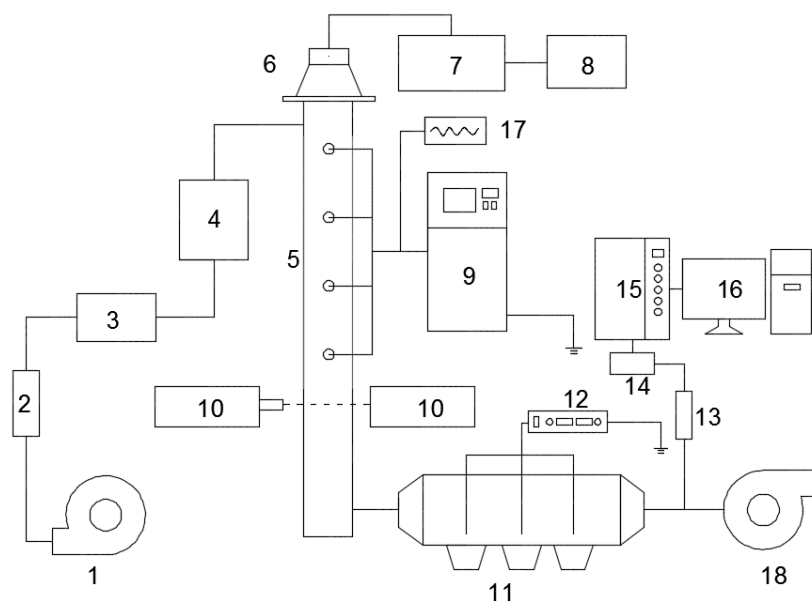
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506   1.Blower 2.flowmeter 3.feeding system 4.buffer tank 5.agglomeration chamber 6.compression

507   driver 7.power amplifier 8.signal generator 9.narrow-pulse high-voltage generator 10.DP-02 laser

508   particle analyzer 11.ESP 12.high voltage DC power supply 13.electrostatic neutralizer 14.diluter

509   15.ELPI 16.computer 17.oscilloscope 18.vacuum pump

510   **Fig. 1.** Schematic of experimental setup.

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Fig. 2. Relationship between the entrainment factor and particle diameter at various acoustic frequencies.
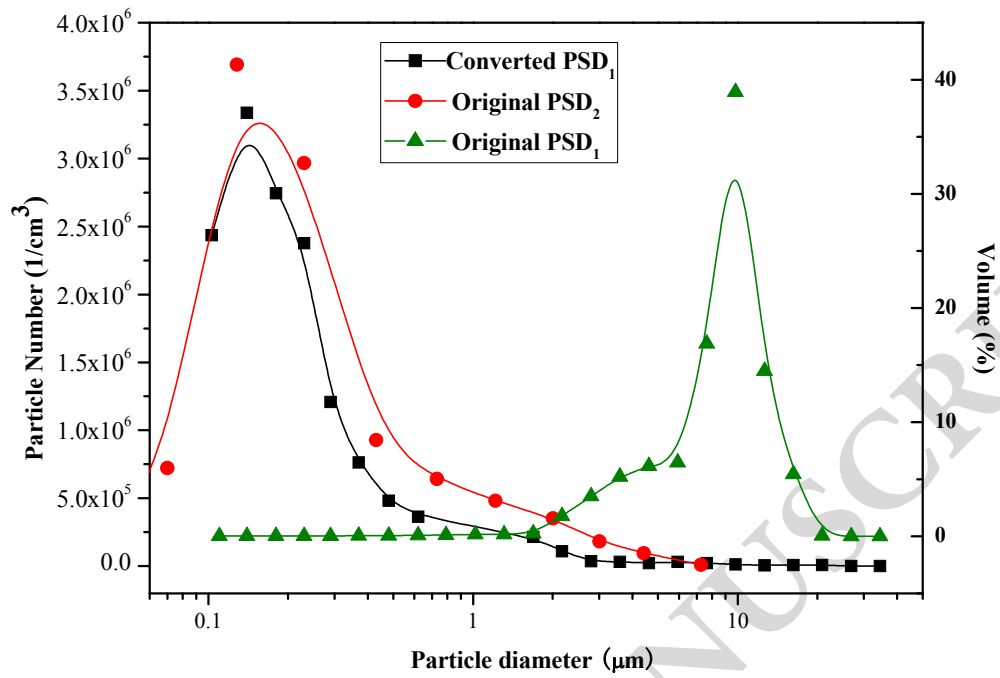
Fig. 3. Original particle size distribution of the ash.

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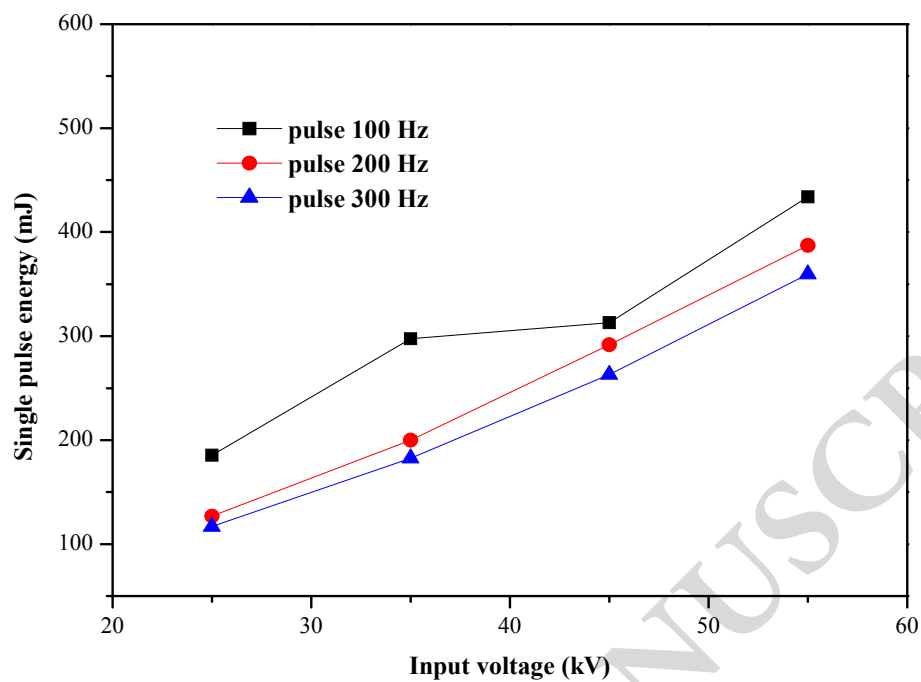
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Fig. 4. Influence of the input voltage on the single pulse energy of pulse corona discharge.

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553 **(a)**

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555 **(b)**

556 **Fig. 5.** Volume differential (a) and cumulative distribution (b) of particle size under various pulse

557 input voltages and frequencies.

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Fig. 6. Median particle diameters and agglomeration ratios of particles under various pulse input

voltages and frequencies.

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(a)

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(b)
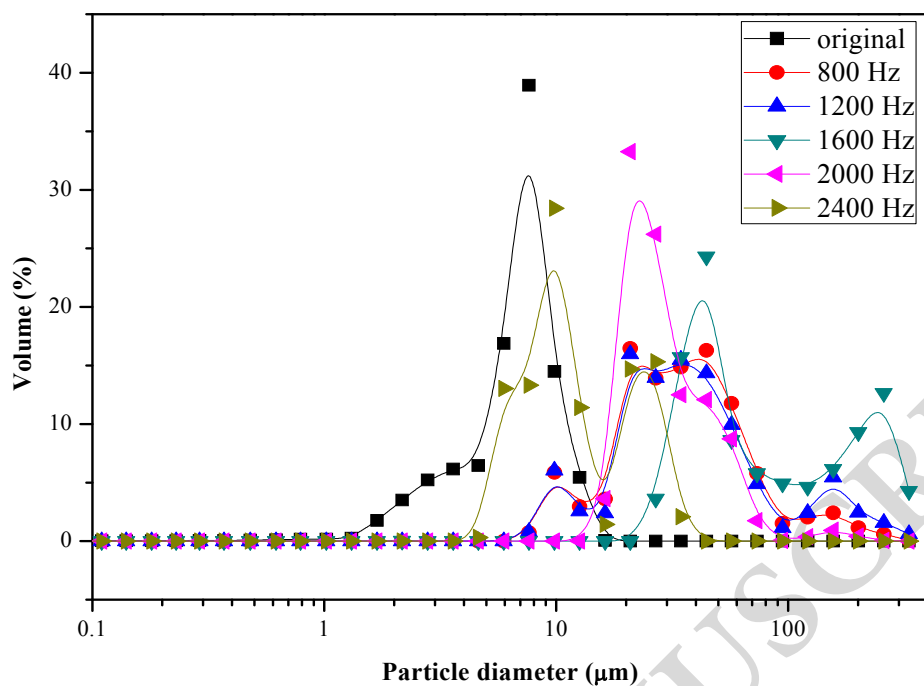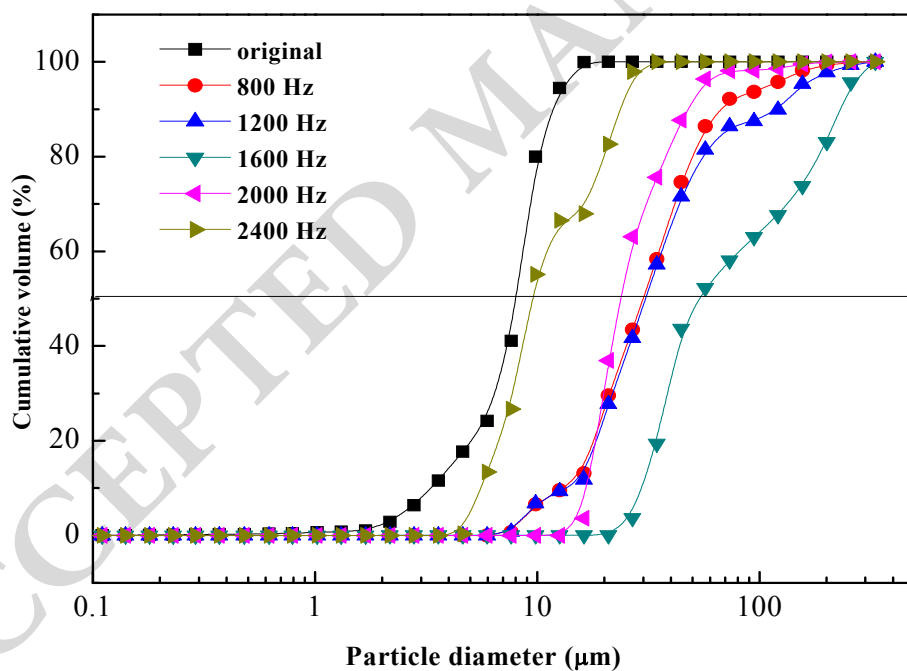
573   **Fig. 7.** Volume differential (a) and cumulative distribution (b) under various acoustic wave

574   frequencies (SPL = 143 dB).

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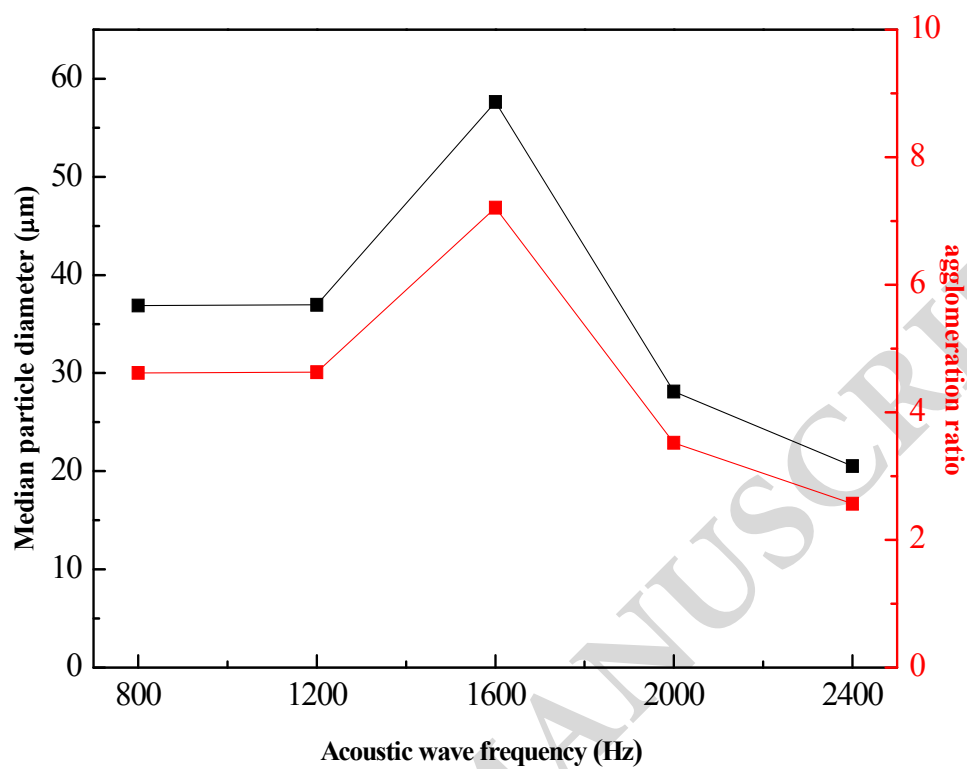
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Fig. 8. Median particle diameters and agglomeration ratios of particles under various acoustic

579   wave frequencies (SPL = 143 dB).

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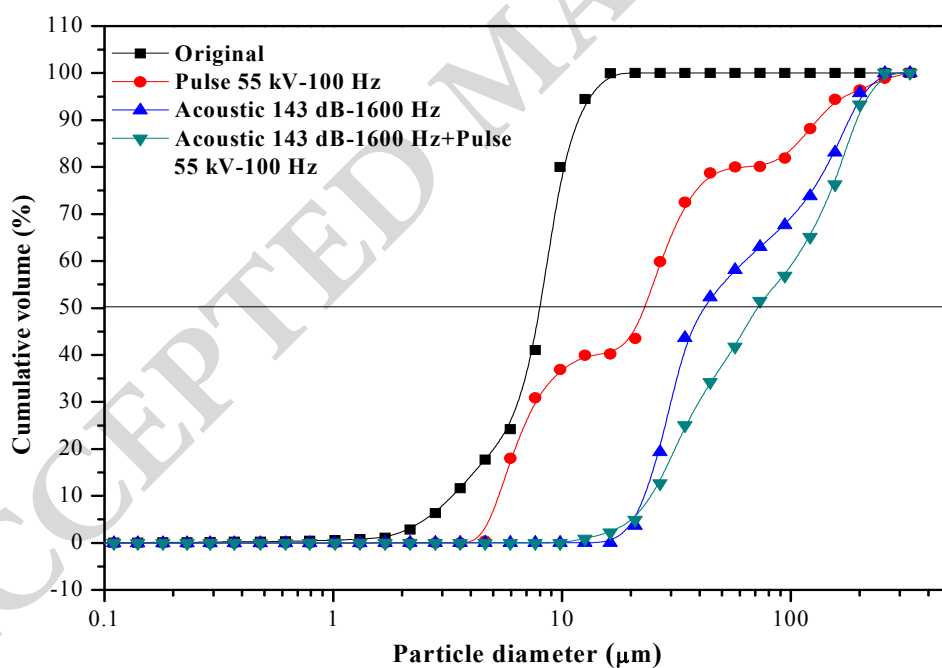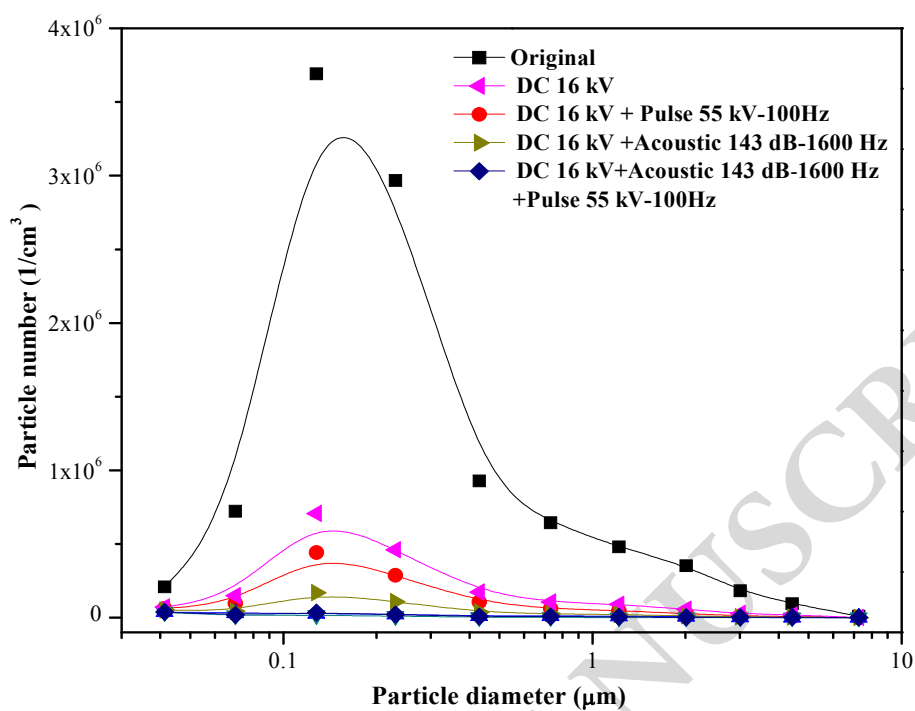
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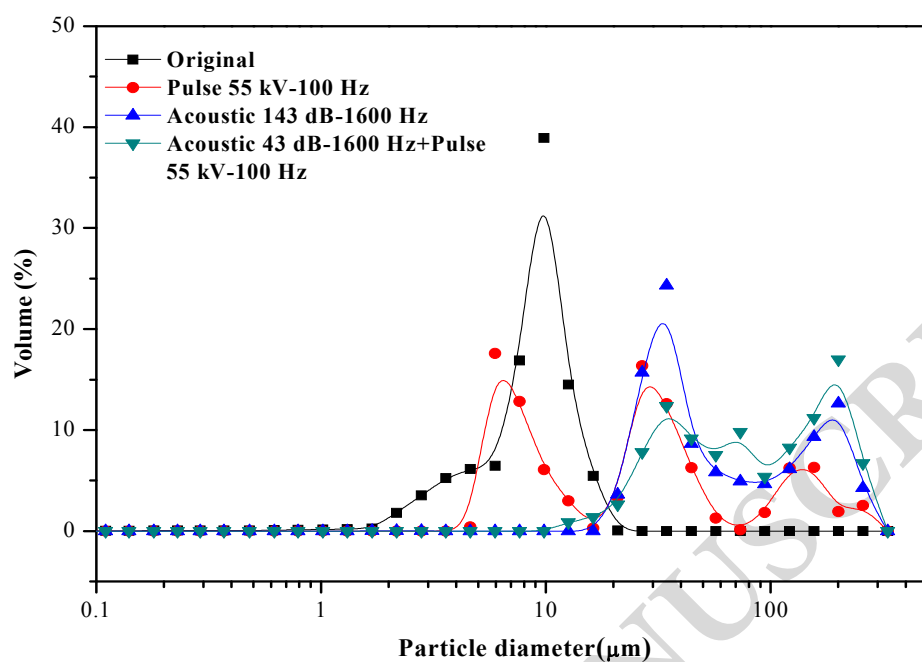
(a)

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(b)

Fig. 9. Volume differential (a) and cumulative distribution (b) of particle size under single field

and coupling field.

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Fig. 10. Comparison of particle number distribution after ESP under various pretreatment

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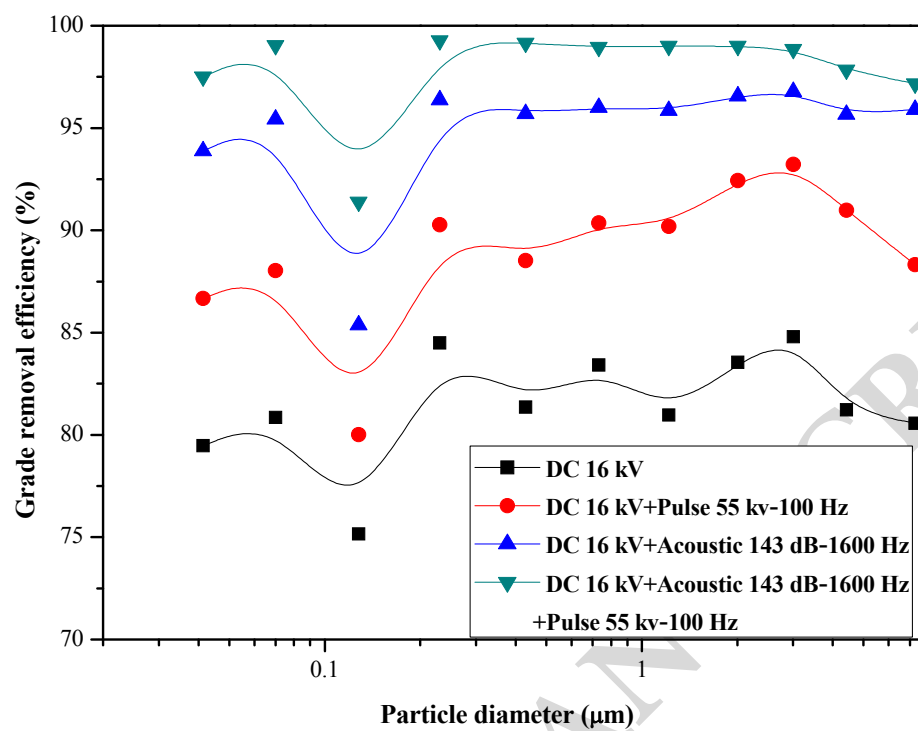599   conditions.

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Fig. 11. Grade removal efficiency of fine particles after application of ESP with single field and

coupling field.

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