Health benefit assessment of China’s National Action Plan on Air Pollution in the Beijing-Tianjin-Hebei area

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Abstract

To evaluate the effect of China’s National Action Plan on Air Pollution (NAPAP), we assessed the health benefits of PM$_{2.5}$ remediation under the NAPAP from 2013 to 2017 in the Beijing-Tianjin-Hebei (BTH) area using a relative risk model with real PM$_{2.5}$ monitoring data and recent statistical research data. The results revealed that the PM$_{2.5}$ concentration in the BTH area decreased by 36 μg m$^{-3}$ (34.0%) under the NAPAP. PM$_{2.5}$-related mortality resulting from all causes, cardiovascular disease, respiratory disease and lung cancer decreased to 58.1-65.2% of that in 2013; 102,133 PM$_{2.5}$-related deaths were avoided, indicating a greater efficacy than the U.S. Cross-State Air Pollution Rule. These results demonstrated that the NAPAP is effective and can be used a reference for other countries to enact similar statutes.

Keywords: Health benefit assessment, PM$_{2.5}$, National Action Plan on Air Pollution, Beijing-Tianjin-Hebei area

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INTRODUCTION

The definition of particulate matter with an aerodynamic diameter less than 2.5 μm (PM$_{2.5}$) was established in 1997 by the U.S. Environmental Protection Agency to protect public health (Liang et al., 2016). Since China began monitoring the ambient air PM$_{2.5}$ concentration nationwide in 2013, PM$_{2.5}$ rapidly replaced PM$_{10}$ as the most common chief pollutant (Pui et al., 2014); the 2013 annual mean PM$_{2.5}$ concentration of the 74 cities required to monitor ambient air PM$_{2.5}$ concentration in phase one (phase two for other cities began in 2016) was 72 μg m$^{-3}$ (ranging from 26-160 μg m$^{-3}$), and 95.9% of the cities failed to meet the secondary annual mean standard (35 μg m$^{-3}$) for ambient air quality in China (Ministry of Ecology and Environment of the People’s Republic of China, 2012).

PM$_{2.5}$ pollution was especially serious in the Beijing-Tianjin-Hebei (BTH) area (Zhang et al., 2017; Cai et al., 2018), where none of the 13 cities met the standard and 7 were listed among the 10 cities with the worst ambient air quality (Ministry of Ecology and Environment of the People’s Republic of China, 2014).

To improve the ambient air quality, as characterized by PM$_{2.5}$, the Chinese government enacted a statute called the National Action Plan on Air Pollution (NAPAP, 2013-2017), which required that the PM$_{2.5}$ concentration in the BTH area be reduced by 25.0% by 2017 compared with that in 2013.
Furthermore, Beijing was asked to reduce the PM$_{2.5}$ concentration to approximately 60 μg m$^{-3}$ (The State Council of the People’s Republic of China, 2013). Many studies assessed the NAPAP by forecasting the health benefits of PM$_{2.5}$ remediation attributed to the NAPAP (Lei et al., 2015; Fang et al., 2016; Chen et al., 2017b; Maji et al., 2018). However, most studies selected the PM$_{2.5}$ concentration target of the NAPAP and utilized outdated population or mortality data as the experimental data. Therefore, the real effect of the NAPAP is still not clear.

The aim of this study is to evaluate the effect of the NAPAP by assessing the health benefits of PM$_{2.5}$ remediation resulting from the NAPAP in the BTH area between 2013 and 2017. The work is based on a relative risk model using real monitoring data and recent statistical research data. We hope the results will be useful for ambient air quality improvement and health risk control.

**MATERIALS AND METHODS**

**Study area**

The BTH area (36°05′N-42°40′N, 113°27′E-119°50′E) is the largest urban agglomeration in northern China and includes Beijing municipality, Tianjin municipality and 11 prefecture-level cities of Hebei Province. In 2016, the area accounted for only 2.2% of the land area but 8.1% of the population and 10.2% of the GDP of China (National Bureau of Statistics of the People’s Republic
of China, 2017). The important position, dense population and poor ambient air quality caused
PM$_{2.5}$-related health problems to be very severe in the BTH area (Zheng et al., 2014; Song et al.,

**Health risk model**

Because the exposure-response relationship between the level of exposure to PM$_{2.5}$ and the
mortality was linear or near linear (Krewski et al., 2000; Pope III, 2000; National Research Council
(U.S.) Committee on Estimating the Health-Risk-Reduction Benefits of Proposed Air Pollution
Regulations, 2002), we selected a widely used epidemiological relative risk model (Broome et al.,
2015; Gao et al., 2016; Chen et al., 2017a) to quantify the association between long-term exposure to
ambient air PM$_{2.5}$ and the risk of PM$_{2.5}$-related disease. The model is as follows:

\[
RR = \exp[\beta \cdot (C - C_0)] \\
I = \left(\frac{RR - 1}{RR}\right) \cdot E \cdot P
\]  

In Eq. (1), $C_0$ represents the critical PM$_{2.5}$ concentration, below which the health impact of PM$_{2.5}$ is
considered to be approximately 0. $C$ represents the exposed PM$_{2.5}$ concentration. $\beta$ is the exposure-
response coefficient, indicating changes in health effects. $RR$ represents the relative risk.

In Eq. (2), $P$ represents the exposed population. $E$ represents the mortality of disease under PM$_{2.5}$
concentration $C$. $I$ represents the PM$_{2.5}$-related deaths, which we used to assess the health benefits of PM$_{2.5}$ remediation resulting from the NAPAP.

The mortalities in this study are from all causes, cardiovascular disease, respiratory disease and lung cancer, as the diameter of PM$_{2.5}$ is small enough to penetrate the bronchioles and alveoli, affecting respiratory function (Turner et al., 2011). Therefore, long-term exposure to PM$_{2.5}$ mainly contributes to the risk of developing cardiovascular disease, respiratory disease and lung cancer (Englert, 2004; Franklin et al., 2007; Kim et al., 2015).

Scenario settings and data sources

In this study, we denoted two scenarios—the baseline and the NAPAP—in the BTH area. The baseline scenario represented the health risk in 2013, whereas the NAPAP scenario represented the health risk under the NAPAP in 2017. The difference in $I$ between the two scenarios was used to assess the health benefits from the NAPAP.

The values of $\beta$ used in this study, given in Table 1, were obtained by meta-analysis. To obtain appropriate values, studies used in the meta-analysis were based on the following criteria: (1) focusing on the long-term association of ambient PM$_{2.5}$ with mortality; (2) including at least one of all-cause mortality, cardiovascular disease mortality, respiratory disease mortality and lung cancer.
mortality; (3) providing effect estimates such as \( \beta \) with standard error or \( RR \) with 95% confidence interval; and (4) studies about China or conducted in recent 3 years were given priority.

The values of \( E \) and \( P \) were mainly obtained from statistical yearbooks, reports and relative studies of each city. For the 2017 data that remain unpublished, we estimated the values of \( E \) and \( P \) from existing data.

\( C_0 \) is defined as the lowest PM\(_{2.5}\) concentration monitored during the exposure time or a threshold recommended by authoritative standards or studies (Apte et al., 2015; Fang et al., 2016). The \( C_0 \) of each city in this study is the air quality guideline (10 \( \mu \)g m\(^{-3}\)) of the WHO (2006), which is a worldwide guideline and close to the primary annual mean standard (15 \( \mu \)g m\(^{-3}\)) for ambient air quality in China (Ministry of Ecology and Environment of the People’s Republic of China, 2012).

The \( C \) under the baseline scenario is 2013 annual mean PM\(_{2.5}\) concentration of each city. The \( C \) under the NAPAP scenario is the PM\(_{2.5}\) concentration of each city under the NAPAP. The annual mean PM\(_{2.5}\) concentration was obtained from the environmental status bulletin of each city.

To assess the health benefits from the NAPAP alone, we excluded the effects of meteorology variations and estimated the PM\(_{2.5}\) concentration under the NAPAP. For the three key cities (Beijing, Tianjin and Shijiazhuang) in the BTH area, the average relative humidity decreased, and the average
wind speed overall increased between 2013 and 2017. Lower average relative humidity results in a weaker secondary formation of PM$_{2.5}$, and higher average wind speed is conducive to the spread of pollutants; thus, the diffusion conditions of the BTH area were considered to be better in 2017 on the whole. In fact, according to official statements (Ministry of Ecology and Environment of the People’s Republic of China, 2018), the meteorological conditions for pollutant diffusion in the BTH area in 2017 were slightly better than those in 2013; the meteorological variations contributed approximately 5.0% to the PM$_{2.5}$ concentration reduction ratio (39.6%), and the NAPAP contributed the remaining 34.6%, accounting for approximately 87.4% (34.6%/39.6%) of the reduction in PM$_{2.5}$ concentration. Thus, the PM$_{2.5}$ concentration under the NAPAP could be calculated based on the 2017 annual mean PM$_{2.5}$ concentration, the PM$_{2.5}$ concentration variation from 2013 to 2017 and the NAPAP contribution.

The details and sources of the above parameters are listed in the Supplementary Material.

RESULTS AND DISCUSSION

PM$_{2.5}$ concentration

We found that under the NAPAP scenario, the PM$_{2.5}$ concentration of each city in the BTH area improved significantly compared with that under the baseline scenario, suggesting that the NAPAP
achieved its target. Fig. 1 shows the PM$_{2.5}$ concentration of each city under the two scenarios. The PM$_{2.5}$ concentration in the BTH area under the NAPAP scenario was 70 μg m$^{-3}$ (ranging from 32-95 μg m$^{-3}$), which is 36 μg m$^{-3}$ (34.0%) lower than that under the baseline scenario (ranging from 40-160 μg m$^{-3}$), thus achieving the target of a 25.0% decrease in PM$_{2.5}$ concentration. The decrease in each city ranged from 20.0-43.8%. Every city except Zhangjiakou and Chengde outperformed the 25.0% target, and Xingtai decreased the most. The low decreases observed in Zhangjiakou and Chengde may be attributed to their already low PM$_{2.5}$ concentrations in the BTH area, which were 40 and 49 μg m$^{-3}$, respectively, under the baseline scenario. Beijing’s PM$_{2.5}$ concentration under the NAPAP scenario was 62 μg m$^{-3}$, and this city also achieved its goal.

PM$_{2.5}$-related mortality

The PM$_{2.5}$-related mortality and PM$_{2.5}$-related mortality ratio (the contribution of the PM$_{2.5}$-related mortality to the overall mortality) in the BTH area under the NAPAP scenario also notably decreased from that under the baseline scenario.

Fig. 2 shows the PM$_{2.5}$-related mortality and the ratio from different causes of deaths in the BTH area under the two scenarios. The PM$_{2.5}$-related all-cause mortality was 1.32‰ under the NAPAP scenario, accounting for 58.1% of that under the baseline scenario. The PM$_{2.5}$-related mortality of
three specific diseases, i.e., cardiovascular disease, respiratory disease and lung cancer, were 0.83, 0.15 and 0.11% respectively, under the NAPAP scenario, accounting for 60.6-65.2% of that under the baseline scenario.

The ratio of PM$_{2.5}$-related all-cause mortality was 23.6% under the NAPAP scenario, 11.0 percentage points lower than that under the baseline scenario. The PM$_{2.5}$-related mortality ratios of cardiovascular disease, respiratory disease and lung cancer were 26.7, 33.4 and 25.0% under the NAPAP scenario, which were 12.1, 13.9 and 11.5 percentage points lower, respectively, than those under the baseline scenario.

Health benefits

We found that the NAPAP had significant health benefits. Fig. 3 shows the deaths and PM$_{2.5}$-related deaths in the BTH area under the two scenarios. For the NAPAP scenario, there were 621,488 deaths, 146,941 (23.6%) of which were related to PM$_{2.5}$. For the baseline scenario, there were 719,084 deaths, 249,074 (34.6%) of which were related to PM$_{2.5}$. Compared with the baseline scenario, 102,133 PM$_{2.5}$-related deaths were avoided in the BTH area due to the NAPAP.

The PM$_{2.5}$-related deaths of the three specific diseases listed above also decreased under the NAPAP scenario, accounting for 60.2-65.4% of that under the baseline scenario. For cardiovascular
disease, there were 344,611 deaths under the NAPAP scenario, 91,985 (26.7%) of which were related to PM$_{2.5}$, which was 58,085 (12.1 percentage points) fewer deaths than that under the baseline scenario. For respiratory disease, there were 49,233 deaths under the NAPAP, 16,448 (33.4%) of which were related to PM$_{2.5}$, which was 8,693 (13.9 percentage points) fewer deaths than that under the baseline scenario. For lung cancer, there were 46,829 deaths under the NAPAP scenario, 11,713 (25.0%) of which were related to PM$_{2.5}$, which was 7,740 (11.4 percentage points) fewer deaths than that under the baseline scenario.

Fig. 4 shows the spatial distribution of PM$_{2.5}$-related deaths in the BTH area under the two scenarios. The average number of PM$_{2.5}$-related deaths in each city in the BTH area under the NAPAP scenario was 11,303 (ranging from 2,479-21,022), 7,856 fewer than that under the baseline scenario (ranging from 4,171-34,718). Shijiazhuang, Beijing and Tangshan were the top 3 most-improved cities, with an average of 13,413 avoided deaths (ranging from 12,149-14,393), while Qinhuangdao, Chengde and Zhangjiakou were the bottom 3 cities, with an average of 1,596 avoided deaths (ranging from 1,476-1,692).

The NAPAP was shown to be more effective than the U.S. Cross-State Air Pollution Rule. The latter rule led to approximately 23,500 avoided deaths per year (Lei et al., 2015), approximated to
117,500 avoided deaths in 5 years for the whole U.S., which is 15,367 (15.0%) more than the avoided deaths in the BTH area under the NAPAP scenario. However, considering the U.S. population was approximately 3 times that of the BTH area, the NAPAP was obviously more efficient.

Economic benefits arising from improved health in the BTH area of the NAPAP were also significant and worthwhile. According to related studies (Zeng and Jiang, 2010; Huang and Zhang, 2013), the value of statistical life in health costs attributable to China’s air pollution was approximately 1 million CNY per person. Therefore, the economic benefits brought about by improved health alone reached approximately 102.1 billion CNY, which accounted for 41.0% of the NAPAP direct investment in the BTH area (Chinese Academy for Environmental Planning, 2015).

**Comparison with related studies**

Previous studies have focused on the same question addressed here. By assuming the exact NAPAP target, (Lei et al., 2015) estimated that the avoided deaths in the BTH area should be 89,000, which was 13,133 (12.9%) fewer deaths than our results. The difference arose from two reasons. The first reason was the difference in $\beta$. The $\beta$ in the previous study, which took the lower limit of that used in the U.S., was 3.0%/10 $\mu$g m$^{-3}$ (ranging from 2.0-4.0%/10 $\mu$g m$^{-3}$), 1.3%/10 $\mu$g m$^{-3}$ lower than
our value. The second reason was the uncertainty in PM$_{2.5}$ concentrations. The PM$_{2.5}$ concentrations in the previous study were from Community Multiscale Air Quality, an air quality forecasting model, which may cause uncertainty in comparison with real monitoring data. Furthermore, the previous study showed a 34.0% decrease in the PM$_{2.5}$ concentrations but provided no detailed figures.

Under the same assumptions, (Chen et al., 2017b) estimated that the deaths prevented in the BTH area should be 70,255 (14,051 per year over 5 years), which was 31,878 (31.2%) fewer deaths than in our study. The difference was mainly attributed to the difference in the PM$_{2.5}$ concentration. The previous study obtained 2012 PM$_{2.5}$ monitoring data from Beijing, Tianjin and Shijiazhuang and selected 2012 as the baseline year. However, the official monitoring and reporting of PM$_{2.5}$ concentrations in China began in 2013, and to the best of our knowledge, no PM$_{2.5}$ concentrations were publicly available in 2012. The 2012 concentration in the previous study was 84 μg m$^{-3}$, which was 22 μg m$^{-3}$ lower than that in our baseline year. Moreover, the concentration in the previous study decreased by approximately 21 μg m$^{-3}$ (25%) from the baseline year to 2017, which was 15 μg m$^{-3}$ lower than the actual result.

According the research results of (Yang et al., 2013), the ambient air PM$_{2.5}$-related mortality ranked fourth in causes of mortality after diet (30.6%, among all reported deaths), high blood
pressure (24.6%), and tobacco (16.4%). Our PM$_{2.5}$-related mortality ratio was 34.6% under the baseline scenario and 23.6% under the NAPAP scenario, both of which are much higher than that of tobacco. However, the study period in the previous study was from 1990 to 2010, and the lack of PM$_{2.5}$ monitoring data and the rapid change in China in the years after may have led to the differences. Nonetheless, the previous study provided a reference; that is, the NAPAP reduced PM$_{2.5}$-related health risks from a level exceeding the risk factor for diet to a level near that of high blood pressure.

Though our results were robust and realistic, some limitations should be noted. Since certain data were unavailable, some of the populations and mortalities were estimated rather than drawn from public statistics. One key index with which to assess the health benefit—the years of life lost (YLL)—could not be obtained, and the exposure-response coefficient was taken from a meta-analysis of the latest study data rather than authoritative results. Thus, we suggest that further studies should focus on local exposure-response coefficients and improving assessments by incorporating the YLL.

**CONCLUSIONS**

We verified the effect of the NAPAP by assessing the health benefits of PM$_{2.5}$ remediation in the
BTH area under the NAPAP. To the best of our knowledge, this is the first and the most accurate study assessing the NAPAP in the BTH area using real PM$_{2.5}$ monitoring data and recent statistical research data.

The NAPAP was effective, as the PM$_{2.5}$ concentration under the NAPAP in the BTH area decreased by 36 $\mu$g m$^{-3}$ (34.0%) from 2013 to 2017, achieving the target set for the area. The PM$_{2.5}$-related mortality caused by all causes, cardiovascular disease, respiratory disease and lung cancer in the BTH area under the NAPAP was 1.32, 0.83, 0.15 and 0.11‰, respectively, accounting for 58.1-65.2% of the values under the baseline scenario. The PM$_{2.5}$-related mortality ratio under the NAPAP was 23.6%, 11.0 percentage points lower than that under the baseline scenario.

Under the NAPAP, 102,133 deaths were avoided in the BTH area and 102.1 billion CNY was saved. The NAPAP was more efficient in improving health than the U.S. Cross-State Air Pollution Rule. Though the PM$_{2.5}$-related mortality in the BTH area decreased from a value exceeding diet-related mortality and reaching the same mortality risk as that of high blood pressure under the NAPAP, the PM$_{2.5}$-related mortality was still high.

This study demonstrated the effect of the NAPAP and proved that the ambient air quality could be improved through effective state planning. We believe that the NAPAP can provide a reference for
countries to improve the ambient air quality and, furthermore, promote the improvement of ambient air quality worldwide.

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DECLARATIONS OF INTEREST

None.

REFERENCES


Table 1. Values of the exposure-response coefficients ($\beta$).

<table>
<thead>
<tr>
<th>Causes of death</th>
<th>$\beta$ (%/10 $\mu$g m$^{-3}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>All causes</td>
<td>4.30</td>
</tr>
<tr>
<td>Cardiovascular disease</td>
<td>4.97</td>
</tr>
<tr>
<td>Respiratory disease</td>
<td>6.69</td>
</tr>
<tr>
<td>Lung cancer</td>
<td>4.66</td>
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</tbody>
</table>
Figure Captions

Fig. 1. PM$_{2.5}$ concentration (μg m$^{-3}$) of each city in the BTH area under the two scenarios

Fig. 2. PM$_{2.5}$-related mortality and PM$_{2.5}$-related mortality ratio in the BTH area under the two scenarios

Fig. 3. Deaths and PM$_{2.5}$-related deaths in the BTH area under the two scenarios

Fig. 4. PM$_{2.5}$-related deaths (thousands) in the BTH area under the two scenarios
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