

Quantification of carbonaceous aerosol emissions from cookstoves in Senegal

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Abstract

In some regions of the world, cooking with solid biomass fuels in open fires constitutes the largest source of elemental and organic carbon emissions. However, cooking-related carbonaceous aerosols are still poorly characterized. This paper presents an innovative characterization of elemental and organic carbon (EC and OC) emissions from cookstoves in West Africa. Four stove types were analysed at laboratory scale (three stones, rocket stove, basic ceramic stove, gasifier), using two wood species (dimb and filao). The EC and OC emission factors based on fuel energy (EFs) when burning dimb were higher for all stoves, highlighting the need to take into account the fuel type when reporting cookstove EFs. The highest EC EF was found with the rocket stove ($0.18 \pm 0.06 \text{ g.MJ}^{-1}$ and $0.06 \pm 0.01 \text{ g.MJ}^{-1}$ for dimb and filao, respectively). The rest of stoves tested showed the same EC EF, when burning dimb ($0.09 \pm 0.02 \text{ g.MJ}^{-1}$) and EC EF ranging between 0.04 ± 0.01 and $0.05 \pm 0.01 \text{ g.MJ}^{-1}$, when burning filao. The average OC EF was $0.08 \pm 0.01 \text{ g.MJ}^{-1}$ for the gasifier, followed by three stones ($0.18 \pm 0.03 \text{ g.MJ}^{-1}$) and the basic ceramic stove ($0.21 \pm 0.08 \text{ g.MJ}^{-1}$). Rocket stove and three stones were also tested under real cooking conditions using wood dimb. Results provide evidence that lab-scale tests overestimated EC EFs measured in the field. Also, the rocket stove didn't show a reduction in wood use with respect to the three stones, implying that carbonaceous aerosol emissions with this stove produce more warming than the traditional stove. Therefore, total stove EC and OC emissions, in addition to EFs, need to be reported. As carbonaceous aerosol impacts are highly dependent on the place where they are emitted, this information can be a very useful input for emission inventories and climate prediction models at national and regional levels.

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41 **INTRODUCTION**

42 In Senegal, 83% percent of households, mainly in rural villages, depend on biomass
43 fuels to cover their daily cooking energy needs(The World Bank, 2013).Combustion of
44 these fuels is usually performed with inefficient stoves, leading to high levels of indoor
45 air pollution and causing the premature death of 6300 people every year, mainly women
46 and children(WHO, 2009).

47 Moreover, cookstove emissions contribute to ambient air pollution and have
48 relevant effects on climate change(Bond, 2004; Chafe et al., 2014; Venkataraman et al.,
49 2005).The carbonaceous components of particulate matter (organic and elemental
50 carbon) emitted during incomplete combustion of biofuels has a strong bearing on the
51 atmospheric radiative balance (Arora and Jain, 2015).

52 Elemental carbon (EC) (or black carbon (BC)) content of particulate matter (PM)
53 has been identified as having the second highest warming impact, despite its short
54 residence time in the atmosphere (Jacobson,2001).BC refers to the light-absorbing
55 carbon, measured by change in light transmittance or reflection, whereas EC is
56 measured by thermal evolution under high-temperature oxidation(Venkataraman et al.,
57 2005). Even though EC and BC are not exactly equivalent (Petzold et al., 2013; Salako,
58 2012), both terms are often used interchangeably (Bond and Bergstrom, 2006; Chen et
59 al., 2005; Li et al., 2009; Shen et al., 2010; Venkataraman et al., 2005).

60 The other major carbonaceous aerosol component, organic carbon (OC), has
61 been linked to light scattering properties (Chung, 2002), although OC from biofuel
62 combustion has been shown to be slightly absorbing due to the presence of brown
63 carbon (Bond et al., 2007; Feng et al., 2013).

64 Aerosols from biomass burning also cause alterations in the environmental
65 conditions at regional and local levels, like visibility reductions and changes in rainfall
66 patterns, which can be substantially large(Ramanathan and Carmichael, 2008). In
67 regions where residential biofuel combustion is prevalent over other energy sources,
68 like Sub-Saharan Africa, Asia and South America, the residential sector is the dominant
69 contributor to BC emissions (Bond, 2004) and worldwide, approximately 25% of BC
70 (although with large uncertainty)and 20% of OC emissions comes from residential
71 burning of biomass fuels(Bond, 2004; Lamarque et al., 2010).

72 However, cooking-related aerosols are much less studied than those of other
73 common sources (Roden et al., 2006; Soneja et al., 2015) like diesel engines, leading to
74 an important underestimation of climate impacts from traditional burning of biomass at
75 residential level (Grieshop et al., 2011; Lee and Chandler, 2013).

76 The complexity of studies of aerosol emissions from cooking is related to the
77 wide distribution and remote location of the sources(Johnson et al., 2008a), the
78 complicated effect of OC(Feng et al., 2013), the relatively complex BC and OC
79 measurement methods (de la Sota et al., 2017; Ramanathan et al., 2011)and because EC
80 and OC emissions range widely between different cooking technologies and biofuels
81 (Grieshop et al., 2011; MacCarty et al., 2008b; Shen et al., 2012; Venkataraman et al.,
82 2005).

83 Some studies that estimate the EC and OC EFs from cooking technologies are
84 available for the Asian Region(Arora and Jain, 2015; Chen et al., 2012; Li et al., 2009;
85 Shen et al., 2013, 2012, 2010; Venkataraman et al., 2005; Zhang et al., 2000); and some
86 contributions have been made from Latin America (Johnson et al., 2008a; Roden et al.,
87 2006, 2009).However, only one study has been performed in sub-Saharan Africa
88 (Malawi) (Wathore et al., 2017) and, to the best of our knowledge, there are no studies

89 regarding the characterization of aerosol emissions from cooking in the West African
90 Region.

91 This paper presents a laboratory scale characterization of aerosol emissions from
92 four stove-fuel combinations widely used in Senegal and other West African countries.
93 Two of these stove-fuel combinations were tested in addition in a rural village of
94 Senegal to understand how real cooking practices in this region influence EC and OC
95 emissions. Finally, total EC and OC emissions and the net climate effect from fuelwood
96 burning at household level in Senegal and West Africa were estimated.

97 A better knowledge of aerosol emissions from residential biofuels in West
98 Africa should contribute to reduce the uncertainties on emission inventories and climate
99 prediction models at regional level, and should result in a more representative and
100 robust estimation of the climatic impacts of this source.

101 Moreover, this information should help improved cookstove initiatives to be
102 considered within the national and regional Short-Lived Climate Pollutant Reduction
103 Strategies, for addressing both near-term and long-term climate change impact, as well
104 as health and social benefits (Shoemaker et al., 2013).

105

106 **METHODS**

107

108 **Laboratory sampling**

109 Laboratory tests were conducted at the Centre de Études et Recherches sur les
110 Énergies Renouvelables (CERER) of the University of Dakar, using a Laboratory
111 Emissions Monitoring System (LEMS), developed by Aprovecho Research Center
112 (ARC) (MacCarty et al., 2010). PM_{2.5} samples were collected via two parallel 102 mm
113 diameter filters, one running at 3 LPM with quartz-fiber filters (Pallflex,
114 tissuquartz,2500 QAT-UP) for OC-EC analyses, and the other running at 16.7 LPM with

115 glass fiber filters (Hi-Q environmental products, FPAE-102) for gravimetric PM_{2.5}
116 measurements. No backup filters were used for quantification of sampling artefacts
117 (Turpin et al., 2000).

118 Three replicates of the Water Boiling Test (WBT) (Bailis et al., 2007) were
119 conducted to test each cooking system. In this standard measurement, 5 litres of water
120 were brought to a boil (cold start phase) and then simmered for 45 minutes. Although
121 the WBT includes three phases (cold start, hot start and simmer), in this study only the
122 cold start and simmer phases were included, to maximise the available resources (time
123 and funds). This simplification can be made when previous WBTs have shown that the
124 cold-start and hot-start phases produce similar results (EPA et al., 2014).

125 The WBT has recognized limitations to reproduce actual random conditions in the
126 field (Johnson et al., 2008b; Roden et al., 2006; Shen et al., 2013). However, its
127 repeatability allows the study of the effect of design on the performance of cookstoves,
128 as well as the emission process influencing factors, mechanisms and kinetics (Arora and
129 Jain, 2015; Chen et al., 2012; Shen et al., 2013).

130 **Field sampling**

131 In-home tests were conducted in March 2016 in Bibane, a rural village in the region
132 of Fatick (Senegal). Emissions testing was done using a Portable Emission Monitoring
133 System (PEMS, ARC), with a similar design as the LEMS, but with a portable hood
134 made from fireproof fabric. The portable hood was placed before starting the meal
135 preparation directly over the centre of the combustion zone (see Figure S1) allowing
136 cooks to have an acceptable space for the stove use. Background concentrations were
137 sampled and then subtracted from the emission sample concentrations.

138 Emission sampling started once the fire was lit, the start time was recorded, and
139 cooks placed the pot on the stove a few seconds after the ignition. At the end of the food

140 preparation stage, emissions sampling was terminated, cooked food weighed, and end
141 time recorded. Fuelwood was measured before and after cooking sessions, as well as the
142 charcoal formed. Our goal was to minimize interference with normal cooking practices,
143 so the food prepared during the tests was freely chosen by families.

144 The PEMS was connected to a diesel generator, placed far enough away to not
145 disturb women while cooking and to avoid emissions interferences.

146 **Stoves and fuels**

147 Four stoves were selected in this study between the most commonly used in rural
148 Senegal, each one representing a method of fuelwood combustion: i) Three-stone (open
149 burning traditional stove);ii) Noflaye Jegg (rocket combustion chamber-type locally
150 produced); iii) Jambaar bois (basic improved ceramic stove, widely distributed in the
151 Dakar region by the German cooperation PERACOD program); iv) Prime Square Stove
152 (a top-lit updraft gasifier ,TLUD, with a natural draft ,with no fan).Photos of the stoves
153 are shown in Figure S2.

154 Each stove was tested in the laboratory using two wood species: Casuarina
155 Equisetifolia (commonly named Filao) and Cordyla Pinnata (commonly named Dimb).
156 Both were split in regular pieces of 15x5x3cm (approximately). Moisture content on a
157 wet basis for dimb and filao was 8% and 8.8%, respectively, and calorific value was
158 21.69 MJ.kg⁻¹ and 19.0 MJ.kg⁻¹, respectively. Three stones and Noflaye Jegg (rocket
159 design) were the stove types analysed in the field. The three-stone fire is still the most
160 commonly used biomass stove in Senegal, and Noflaye Jegg stove has been widely
161 implemented in Senegal and neighbouring countries(Sota et al., 2014).

162 In 2012, within the framework of a program to optimize the use of forest resources
163 in the region, 50 Noflaye Jegg stoves were distributed in the Bibane. However, only 15
164 out of 50 stoves were in good or fair state, and considered adequate to be included in the

165 study and some of them needed minor repairs before starting the tests. Finally, 15
166 households with improved cookstoves were included in the study, together with another
167 14 using the traditional three stone fire.

168 As fuel used in Bibane was quite heterogeneous due to fuelwood scarcity, the
169 same wood was distributed to every family participating in the study. Testing with
170 uniform wood type allowed us to reduce the variability between households and to
171 focus the study on the effect of stove type on carbonaceous aerosol emissions. Wood
172 used during field testing (dimb) was purchased from a local vendor, bigger and more
173 irregular pieces than those used in the laboratory, and distributed to households.
174 Purchased dimb had a moisture content of 5% on a wet basis and a calorific value for
175 dry fuel of 21.69MJ.kg⁻¹.

176 **EC and OC analysis**

177 After sampling, particle-loaded filters were packed with aluminium foil and
178 stored in a freezer until analysis. Gravimetric measurements of glass fibre filters were
179 conducted using a high precision digital balance at the CERER, and OC/EC analyses on
180 quartz filters were conducted with a Sunset Laboratory Carbon Analyser using the
181 EUSAAR2Protocol(Cavalli et al., 2010),in the Institute of Environmental Assessment
182 and Water Research IDAEA-CSIC in Barcelona. This protocol has been already used in
183 cookstove aerosol emission analysis (de la Sota et al., 2017).

184 **Data Analysis**

185 Kruskal-Wallis one-way analysis of variance was used to identify significant
186 differences of OC and EC EFs with different cookstoves, fuels and during different
187 WBT phase. Wilcoxon t-test was conducted in order to identify pairs with statistically
188 significant difference in EF values. Non parametric tests were used because

189 observations were not normally distributed, due to the relative small sample size and the
190 presence of outliers (mainly in results from the field study).

191 The inter quartile range (IQR), difference between the 75th percentile and the
192 25th percentile, was used to determine the test-to-test variability. This measure is robust
193 against outliers and non-normal data. All statistical analyses were conducted at a
194 significance level of 0.05 and performed using IBM SPSS Statistics Base software.

195 **RESULTS AND DISCUSSION**

196

197 **Laboratory estimation of EC and OC emission using the WBT**

198 Figure 1 presents EC EFs (in grams per MJ fuel energy, for comparison on the
199 same scale), total fuel consumption per WBT and EC total emissions per WBT
200 completed (cold start + simmer).

201 Results on Figure 1a show that the type of stove had an effect on EC EFs. The
202 highest average EC EF per WBT was found with the Noflaye Jegg for both fuels (dimb
203 and filao), while EC EF's were fairly uniform across the other two improved cookstoves
204 and the three stones.

205 Differences of EC EF's between stoves are related to the nature of combustion
206 (Arora and Jain, 2015; MacCarty et al., 2007), which is, in turn, affected by a number of
207 factors (e.g. air-fuel ratio, burn rate, combustion temperature, combustion efficiency,
208 thermal efficiency, residence time in the combustion chamber, flame turbulence, etc.)
209 (MacCarty et al., 2008b; Venkataraman et al., 2005).

210 In this study, the CO/CO₂ ratio was used as a benchmark for stove combustion
211 efficiency(Li et al., 2009; Shen et al., 2012).The Noflaye Jegg (rocket stove) performed
212 with the lowest CO/CO₂ ratios (3.4±1.1% for dimb and 5.3±0.1% for filao), indicating

213 an acceptable efficient combustion (target value is 2% or less), and the highest EC EFs
214 ($0.18\pm 0.06 \text{ g.MJ}^{-1}$ and $0.06\pm 0.01 \text{ g.MJ}^{-1}$) for dimb and filao, respectively.

215 Rocket stoves have a strong draft which enhances air flow through the fire,
216 improving the combustion efficiency and resulting in higher combustion temperatures
217 (Li et al., 2009; MacCarty et al., 2008b), which typically results in higher EC
218 emissions (Bond, 2004; Cachier et al., 1996; MacCarty et al., 2008a; Rau, 1989).

219 The three stones showed the highest CO/CO₂ ratios ($8.7\pm 0.5\%$ with dimb and
220 $8.3\pm 1.8\%$ with filao), indicating inefficient combustion, and low values of EC EF
221 ($0.09\pm 0.01 \text{ g.MJ}^{-1}$ for dimb and $0.04\pm 0.01 \text{ g.MJ}^{-1}$ for filao).

222 Jambaar bois showed very similar values to the three stones fire, both in
223 combustion efficiency and EC EFs values. CO/CO₂ ratio was $7.8\pm 2.4\%$ and $6.7\pm 1.9\%$;
224 and EC EFs were $0.09\pm 0.01 \text{ g.MJ}^{-1}$ and $0.05\pm 0.01 \text{ g.MJ}^{-1}$, with dimb and filao,
225 respectively. Previous studies have also found EFs for basic ceramic stoves similar to
226 traditional stoves (Wathore et al., 2017).

227 The gasifier showed CO/CO₂ ratio of $3.6\pm 0.6\%$ and $4.0\pm 0.4\%$, with dimb and
228 filao, respectively (i.e., high combustion efficiency); and low EC EF's $0.09\pm 0.02 \text{ g.MJ}^{-1}$
229 when burning dimb, (equal to EF of Jambaar Bois and three stones) and 0.05 ± 0.01
230 g.MJ^{-1} with filao. This is coherent with previous findings showing the lowest EF for
231 gasifiers in comparison with traditional and other basic improved stoves (Arora and Jain,
232 2015; MacCarty et al., 2008b; Wathore et al., 2017), (Table S1).

233 In addition to the cookstove technology, the biofuel characteristics also
234 determine combustion conditions, and thus, aerosol emissions (Li et al., 2009;
235 Venkataraman et al., 2005; Venkataraman and Rao, 2001). In our study, EC EF's of all
236 the stoves tested showed higher values when burning dimb over filao (64.2%, 51.7%,

237 49.7% and 56.7% for the Noflaye Jegg, Jambaar bois, gasifier and three-stones,
238 respectively), although not statistically significant ($p>0.05$).

239 A number of characteristics of fuelwood could affect pollutant emissions: i) the
240 geometrical shape and size of the fuel; ii) the density (Atiku et al., 2016; Mitchell et al.,
241 2016); iii) the lignin content (Atiku et al., 2016); iv) the heat value and v) the moisture
242 (Chomanee, 2009; Shen et al., 2010), among others.

243 Dimb and filao used in this study had similar moisture fraction, calorific value
244 and size. Typical lignin content is around 26% for Filao (Mahmood, 1993), and 40% for
245 Dimb(Samba, 2001).Previous studies show that lignin content promotes elemental
246 carbon formation (Atiku et al., 2016; Tillman, 1987), so this may partially explain the
247 higher EC EF's of wood dimb.

248 Results suggest that emission factors of EC are dependent on the type of fuel
249 burned, but further studies could examine the effect of physical and chemical
250 characteristics of filao and dimb in EC EF's, which cannot be explained at this stage.

251 Figure 1b presents EC EFs for the stove-fuel combinations by WBT phase,
252 except in the case of the gasifier, where only the EF of the whole WBT is presented,
253 because its cylindrical combustion chamber cannot be emptied during the burning
254 cycle.EC EFs were found to be higher during the cold start phase for every stove-fuel
255 combination analysed. This can be explained because cold start phase is characterized
256 by strong flames and significant amounts of elemental carbon (Atiku et al., 2016; Roden
257 et al., 2009), especially shortly after fire ignition (Roden et al., 2006), whereas the
258 simmer phase is characterized by smaller flames, with lower emission of EC particles
259 (Roden et al., 2006).

260 EC EFs were 20%, 10% and 29% higher for the cold start phase with regard to
261 simmer, for the dimb with three stone, Jambaar Bois and Noflaye Jegg, respectively;

262 and 25%, 20% and 14% higher for filao with three stone, Jambaar Bois and Noflaye
263 Jegg, respectively; though not statistically significant ($p>0.05$). Previous studies also
264 found that changes in combustion conditions during the cooking cycle induced variation
265 in EC EFs (Arora and Jain, 2015; Jetter et al., 2012; Roden et al., 2006).

266 In terms of total EC emission per WBT completed, Figure 1d shows that results
267 followed almost the same trend as in the case of average EC EF's per WBT. Noflaye
268 Jegg showed the highest EC EF and didn't show a reduction in fuel wood consumption
269 per task completed with respect to the three stones fire (Figure 1c). Therefore, this type
270 of stove has the highest EC emissions per WBT (3.75 g per WBT with dimb and 1.90 g
271 per WBT with filao). On the other hand, the gasifier showed the highest fuel savings and
272 low EC EFs, so this type of stove achieved the smallest total EC emissions per WBT
273 (1.35 g per WBT with dimb and 0.57 g per WBT with filao). An intermediate situation
274 was found for the three stones and the Jambaar bois, which show very similar results in
275 terms of EC total emissions (three stones, 2.00 g per WBT with dimb and 0.67 g per
276 WBT with filao) and Jambaar bois (1.94 g per WBT with dimb and 0.89 g per WBT
277 with filao). The fuel consumption value obtained for the Noflaye Jegg with filao (and,
278 therefore, the value of total emissions) in this study should be considered with caution,
279 since in the other stove types tested, fuelwood consumption used does not vary with the
280 species used.

281 Figure 2 presents OC EFs for three stones, Jambaar Bois, and gasifier, when
282 burning filao. OC EFs for the rest of stove-fuel combinations are not presented due to
283 problems in filter handling between WBT and EC/OC analysis.

284 When burning filao wood, the average OC EF for the whole WBT was found to
285 be the lowest for the gasifier (0.08 ± 0.01 g.MJ⁻¹), followed by three stones (0.18 ± 0.03
286 g.MJ⁻¹) and then Jambaar bois (0.21 ± 0.08 g.MJ⁻¹). Low OC and EC emissions of gasifier

287 are explained because this technology emits very low levels of particulate matter when
288 compared to other biomass stoves (MacCarty et al., 2008b). As already observed in
289 previous studies, in the case of the three-stone fire and Jambaar bois, a large bed of
290 charcoal under the flaming fuel resulted in smoldering conditions, with high OC
291 particles emissions (MacCarty et al., 2007). Overall, our OC EF results appear to be
292 consistent with prior studies (Table S1).

293 Test phases that were associated with higher OC emissions had different results
294 depending on the cookstove tested, as occurred in previous studies (Arora and Jain,
295 2015). For the traditional stove, OC EF was found to be higher during simmer
296 ($0.19 \pm 0.05 \text{ g.MJ}^{-1}$) than during cold start ($0.16 \pm 0.03 \text{ g.MJ}^{-1}$), whereas Jambaar bois
297 showed highest OC EF during cold start ($0.22 \pm 0.09 \text{ g.MJ}^{-1}$) when compared to simmer
298 phase ($0.19 \pm 0.05 \text{ g.MJ}^{-1}$), although none of these differences between WBT phases were
299 statistically significant (Wilcoxon, $p < 0.05$).

300 Akagi et al., 2010, attributed higher OC emissions during the simmer phase to
301 high carbon content and large surface area per unit mass of charcoal that results in the
302 formation of organic compounds. On the other hand, higher OC emissions during
303 flaming conditions (cold start phase) with the Jambaar Bois, could be attributed to low
304 flame temperatures due to heat losses through the cookstove walls (Arora and Jain,
305 2015).

306 Figure 2d presents results of OC/EC ratio, considered an environmentally
307 relevant property with good predictive capabilities in determining the impact of
308 absorbing aerosol on the Earth's radiative balance, even for fuels in which brown
309 carbon absorption is significant (Pokhrel et al., 2016). Overall, all stove fuel
310 combinations showed OC/EC ratios higher than one, indicating the dominance of OC
311 and suggesting that the smoldering combustion was dominant during the WBT.

312 However, real-time data such as the modified combustion efficiency (MCE;
313 $\text{CO}_2/(\text{CO}+\text{CO}_2)$) and Single Scattering Albedo would be better proxies to differentiate
314 between smoldering and flaming dominated combustion.

315 Figures 1 and 2 show a relatively large degree of test-to-test variability (boxes
316 overlapped considerably). For this reason, differences in OC and EC EFs across stove
317 types, phases of WBT and types of fuel were not found to be statistically significant in
318 most cases ($p>0.05$).

319 In the Laboratory evaluation of EC emissions using the WBT (Figure 1), the
320 highest test-to-test variability was observed with the Noflaye Jegg (inter quartile range,
321 $\text{IQR}=2.8$ with dimb and $\text{IQR}= 1.03$ with filao, figure 1d). This is because flaming
322 conditions are more present in the cooking cycle for this stove, characterized by more
323 frequent peaks of emissions, which significantly affect total EC. For OC emissions,
324 Jambaar bois showed the highest IQR, equal to 2.9 (figure 2c).

325 In this study three replicates of WBT were conducted because it was commonly
326 used in the literature (Wang et al., 2014). However, number of replicates is linked to the
327 stove-fuel combination being tested and sometimes three are not enough to ensure
328 confidence in the results (Wang et al., 2014). Our findings suggest that in forthcoming
329 emissions testing of the stoves included in this study, a larger number of WBT
330 replicates, especially for the stoves with greater variation in performance, should be
331 analysed.

332 **Field estimation of EC and OC emission factors**

333 Most families cooked inside the kitchen space during the tests (separated from
334 the rest of the house and very poorly ventilated), although a few of them cooked
335 outdoors. They prepared typical meals of rural households in the region, that consisted
336 of rice with vegetables, fish, or meat for lunch, whereas for dinner, water was boiled to

337 steam cous-cous and a sauce was prepared. Time of preparation ranged from 38 to 111
338 minutes (average=75 minutes), with no significant differences between lunch and dinner
339 and between the type of stove used. Likewise, no differences were found in the quantity
340 of fuel used per kg of food prepared between the three stones and the Noflaye Jegg (0.6
341 kg of fuel per kg of food), consistent with previous laboratory results in this study.

342 As fuelwood type was the same for every household, our field results focused on
343 the analysis of the influence of stove type on EC and OC emissions. , EC EFs (Figure
344 3a), increased a 75% ($p < 0.05$) with Noflaye Jegg in comparison with the three stones
345 stove. On the other hand, OC EF's (Figure 3b) were 30.3% and 35.7% lower for Noflaye
346 Jegg when compared to the three stones, though not statistically significant ($p > 0.05$).
347 The ratio OC/EC (Figure 3c) was 1.77 ± 0.56 for the three stones, indicating the
348 dominance of OC, which may be caused by the smoldering combustion (MacCarty et al.,
349 2008a). OC/EC ratio < 1 (0.67 ± 0.28) for the Noflaye Jegg is showing the dominance of
350 EC produced by flaming conditions, typical in the rocket stoves (MacCarty et al.,
351 2008a).

352 Therefore, EC EFs are higher with the Noflaye Jegg and thus this stove may be
353 more warming (from a climate perspective, considering only EC and OC) when
354 compared to the three stones, which could lessen some of the impact of reducing overall
355 aerosol emissions (Johnson et al., 2008a). However, the comparison of both stoves
356 should be done in terms of total EC and OC emissions, taking into account their fuel-
357 efficiency (see section *Estimation of total emissions and net climate effect of cooking-*
358 *related carbonaceous aerosol in Senegal*).

359 Overall, EC and OC EFs obtained agree well with previous studies (Table S1),
360 although many differences were observed due to differences on stove type, fuel
361 parameters, operation conditions, etc. Nonetheless, most publications present EF in

362 g.kg⁻¹ instead of g. MJ⁻¹, which prevents comparison between different fuel types at
363 same level.

364 **Laboratory vs field results**

365 EC EFs from WBT were compared with results obtained under field conditions
366 for the three stones and the Noflaye Jegg stoves when burning dimb (Figure 4).

367 EC EFs estimated at laboratory scale were 18.0% higher than during daily
368 cooking activities for the NoflayeJegg, and 14.5% higher in the case of the three stones
369 (not significant, $p>0.05$). These results suggest an overestimation of EC EFs at
370 laboratory scale, when compared with field conditions. Previous studies also found
371 differences between results from laboratory tests and those obtained during field
372 conditions (Arora and Jain, 2015; Chen et al., 2012; Johnson et al., 2008b; Roden et al.,
373 2009, 2006; Shen et al., 2013), although differences were larger than in this study.

374 Differences between laboratory and field results can be explained by several
375 factors. First, the fire management behaviour in the laboratory was essentially
376 normalized, the cooking fire was constantly tended and fuel was added frequently,
377 whereas the process observed in the field was rather random, including some low
378 combustion efficiency situations due to infrequent tending. Wood pieces used in the
379 laboratory had regular shape and dimensions, while in the field they were less uniform
380 in shape and with larger dimensions. Smaller wood has been found to yield lower PM
381 emissions and a higher BC fraction of emissions (Bond, 2004; Li et al., 2009). Thirdly,
382 stove age and may affect the stove fuel consumption and emissions (Naehler et al., 2000).
383 Finally, minor fugitive emissions may have escaped from the fabric hood during field
384 testing.

385 Differences between lab and field EFs were lower during the simmer phase of
386 WBT: 3.8% for the three stones and 4.2% for the Noflaye Jegg ($p>0.05$) (see Table S4,

387 supporting information). This is coherent with our field observations, which showed that
388 during meal preparations, cooks maintained a constant and low-medium flame during
389 almost all the time, except when frying for the first few minutes, which needed stronger
390 flames.

391 Although user practices are highly individual, regional averages may be
392 governed by common customs (Chen et al., 2012), simmer phase of WBT is likely quite
393 illustrative of EC emissions in homes of rural villages of Senegal. Further field
394 measurements are needed as a reality check for each stove-testing and region.

395 **Estimation of total emissions and net climate effect of cooking-related** 396 **carbonaceous aerosol in Senegal**

397 Total annual EC and OC emission per year of use of the Noflaye Jegg and the
398 three stones were calculated with EFs values and data of daily consumption (9.9 ± 5.1 kg
399 per day for the Noflaye Jegg and 9.6 ± 5.5 kg per day for the three stones), both
400 determined during the field study. As in this region cooking activities occur most often
401 in indoor environments, it was necessary to take into account the fraction of emissions
402 released to ambient air (exfiltration rate), which is dependent on the ventilation and
403 particle deposition rates in kitchens (Soneja et al., 2015). Exfiltration rate for typical
404 household characteristics of this study was not available, so total emissions have been
405 estimated using two values of exfiltration rate from previous studies (26% and
406 80%) (Soneja et al., 2015 and Venkataraman et al., 2005, respectively). Annual EC and
407 OC emissions per stove were then factored by the Global Warming Potential (GWP)
408 value recommended by the IPCC, 900 for EC (Bond et al., 2013) and -46 for OC (Bond
409 et al., 2011), for a time horizon of 100 years.

410 Figure 5 presents the net climatic effect estimated with regard to EC and OC
411 emissions for the three stones and Noflaye Jegg (in tons of $\text{CO}_2\text{-eq}$ per year). At this

412 point, it is important to highlight that a complete assessment of the global warming
413 potential should include other species co-emitted from residential cookstove use
414 (Bhattacharya et al., 2002; Roden et al., 2006; Smith et al., 2000). Using the data of fuel
415 consumption and emission factors obtained in this study, across a 100-year horizon,
416 Noflaye Jegg showed an overall increase of 88% global warming potential in
417 comparison to the traditional stove. In consequence, while the use of Noflaye Jegg has
418 clear benefits for improving indoor air quality (Sota et al., 2014), our results suggest
419 that this type of rocket stove, under typical cooking practices and fuelwood used in
420 Senegal, may lead to negative climate effects with regard to carbonaceous aerosol
421 emissions, although a complete analysis would include co-emitted species.

422 The fact that rocket and other types of natural draft stoves increase EC EFs is
423 known (Arora and Jain, 2015; MacCarty et al., 2008a). The unexpected point is that
424 Noflaye Jegg stove did not provide fuel saving benefits, neither in our laboratory study,
425 nor in the field. For this reason, our work emphasizes the importance of taking into
426 account total EC and OC emissions in addition to EFs.

427 Besides this, total annual emissions of EC and OC from residential burning of
428 fuel wood in Senegal were estimated using the average value of EC and OC EFs
429 determined in this study ($1.85 \text{ g EC.kg}^{-1}$ and $3.97 \text{ g OC.kg}^{-1}$) and the total annual
430 firewood used in Senegalese households (450 ktOE per year, for the period 2000-
431 2005)(Dafrallah, 2009).

432 As previously discussed, EC and OC EFs vary significantly across different
433 types of stove-fuel combinations. However, the average EF's from this study could be
434 considered as representative of Senegal, as they were estimated from a set of biomass
435 cookstoves and wood species widely used in the country. Fuel consumption is also
436 highly dependent on type of stove used, but data on the fractions of different stoves

437 used in Senegal are unavailable, so the total firewood annually used in the Senegalese
438 households was the best information available.

439 Total emissions of EC and OC from residential fuelwood combustion in Senegal
440 were estimated to be 421.1 ± 213.0 and 561.4 ± 411.4 t per year, respectively, with 26% of
441 exfiltration rate, and 1295.6 ± 655.3 t per year and 1727.5 ± 1265.8 t per year with 80% of
442 exfiltration rate. Factoring by GWP₁₀₀ values, net effect of EC and OC emissions from
443 household fuelwood consumption in Senegal were 353.1 kt of CO₂-eq with 26% of
444 exfiltration rate and 1086.6 kt of CO₂-eq with 80 % of exfiltration (Figure 6).

445 Using the 5th percentile of the EC and OC EFs data set and 26% of exfiltration
446 rate, the result was 173.4 kt of CO₂-eq, and, when using the 95th percentile of EC and
447 OC EF's and 80% of exfiltration rate, calculations show a result of 2031.9 kt of CO₂-
448 eq. This range (173.4-2031.9 kt of CO₂-eq) constitutes a rough estimation of the net
449 effect on climate of carbonaceous aerosols emitted into the atmosphere from household
450 use of fuelwood for cooking in Senegal.

451 It is important to highlight that a complete assessment of the global warming
452 potential should include other species co-emitted from residential cookstove use
453 (Bhattacharya et al., 2002; Roden et al., 2006; Smith et al., 2000). However, the
454 objective at this stage was to assess the net climate effect of cooking-related
455 carbonaceous aerosol emissions in Senegal, which is still uncertain.

456

457 **CONCLUSIONS**

458 This work provides original cooking-related carbonaceous aerosols emissions
459 estimates in Senegal, where data were not available. This information can provide a
460 very useful input for atmospheric models and for supporting future mitigation strategies
461 at national and regional level.

462 However, uncertainties in EF's estimates remain, as well as in the quantities of
463 fuel consumed. Moreover, the 100-year GWP value for black carbon ranges from
464 120 to 1800, when all forcing mechanisms are included, and can vary by about $\pm 30\%$
465 between emitting regions(Henze et al., 2012; Lacey and Henze, 2015; Myhre et al.,
466 2013).

467 Further laboratory and field studies should be performed in other regions of
468 Senegal and West Africa with different stove-fuel combinations, also studying the
469 seasonality effect on carbonaceous emissions and considering the warming effect of
470 brown carbon.

471 These studies should be especially focused on biomass-based advanced
472 technologies and fuels, such as gasifiers and pellets, as they will most likely be the more
473 accessible and affordable clean cooking solutions in this region in the short-medium
474 term. To the extent possible, EFs should be characterised in the field and under real-
475 world cooking conditions.

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482

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693 **FIGURE CAPTIONS**

694

695 -Figure 1. Laboratory estimation of EC emission using the WBT. Figure 1a: EC EFs per
696 MJ fuel energy for the WBT; Figure 1b: EC EFs per MJ fuel energy for each WBT
697 phase (cold-start and simmer); Figure 1c: Fuel consumption per WBT; Figure 1d: Total
698 EC Emissions per WBT.

699

700 -Figure 2. Laboratory estimation of OC emission using the WBT. Figure 2a: OC EFs
701 per MJ fuel energy for the WBT; Figure 2b: OC EFs per MJ fuel energy for each WBT
702 phase (cold-start and simmer); Figure 2c: Total OC emissions per WBT; Figure 2d:
703 OC/EC ratio.

704

705 - Figure 3. Field estimation of EC and OC emissions. Figure 3a: EC EFs; Figure 3b: OC
706 EFs; Figure 3c: OC/EC.

707

708 - Figure 4. Boxplot of EC EF (g/MJ) for each type of stove and test.

709

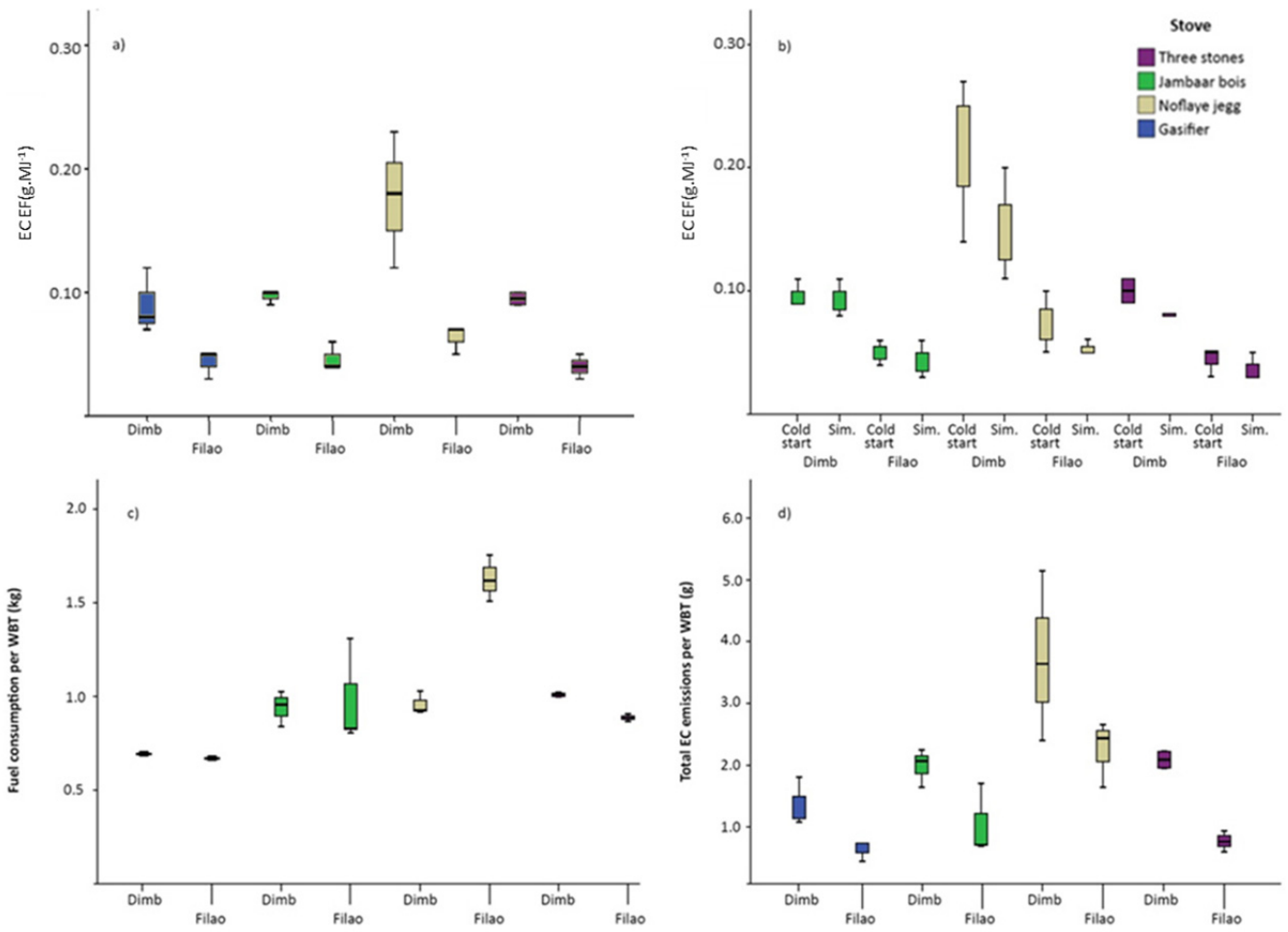
710 - Figure 5. tonnes of CO₂-equivalent per year per stove (calculated with 100-yr GWP)
711 from EC and OC emitted during one year of use of NoflayeJeg stove and Three stones.
712 (Co-emitted species are not included in the analysis).

713 - Figure 6. tonnes of CO₂-equivalent per year in Senegal(calculated with 100-yr
714 GWP)from EC and OC emitted from household fuelwood cookstoves, for an
715 exfiltration rate of 26% and 80%.

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718 Fig. 1.



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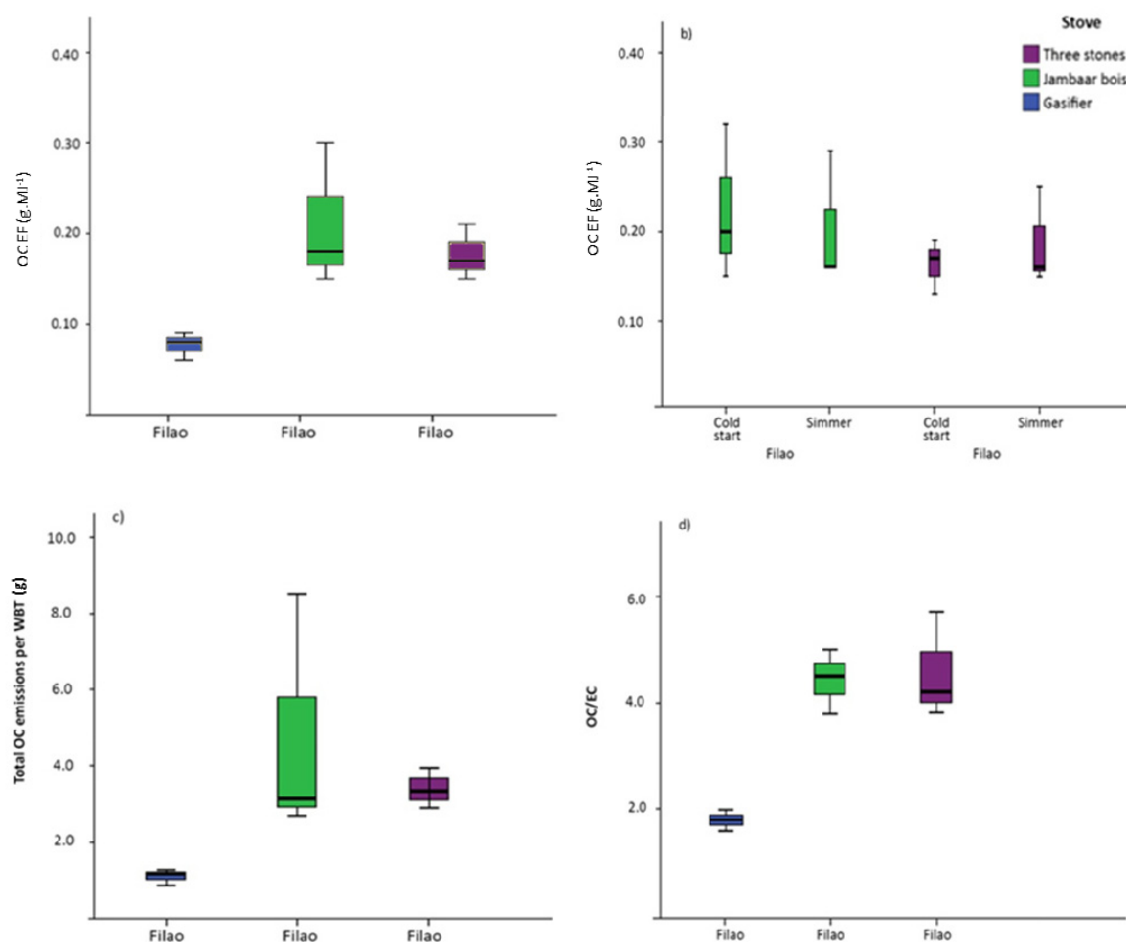
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721 Fig. 2.

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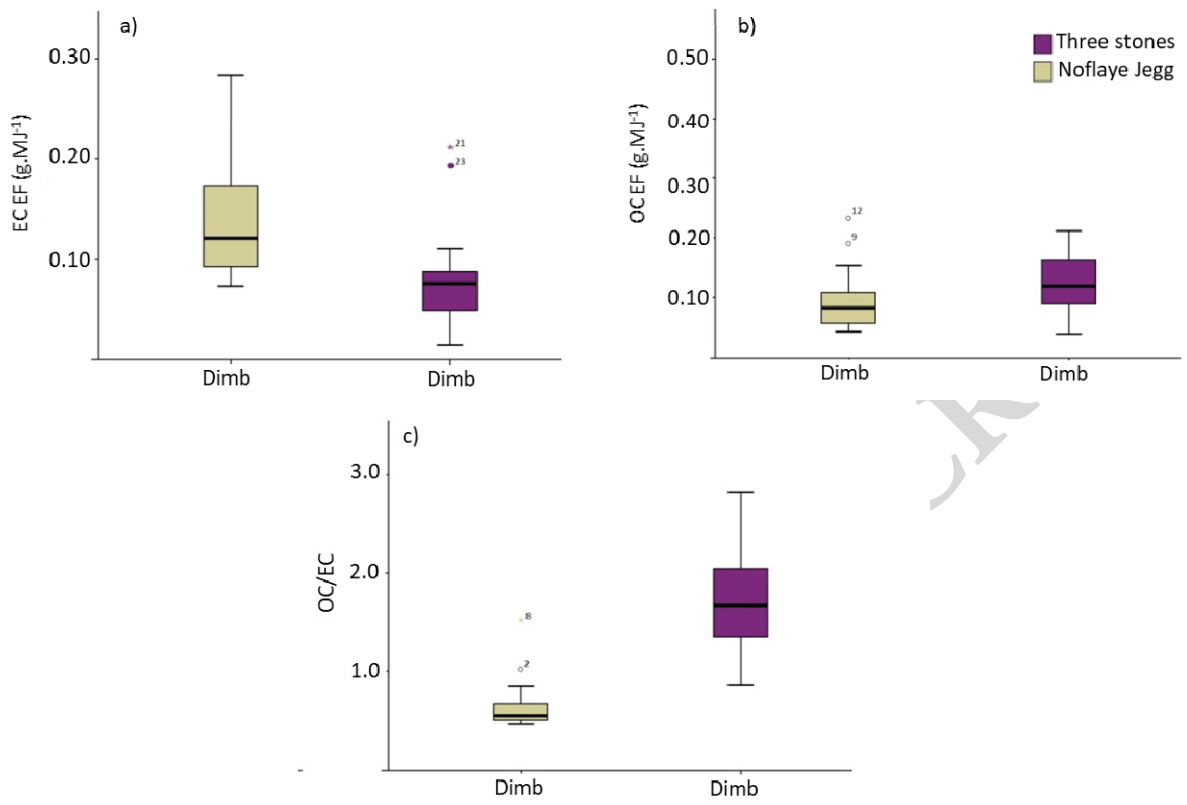


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724 Fig.3

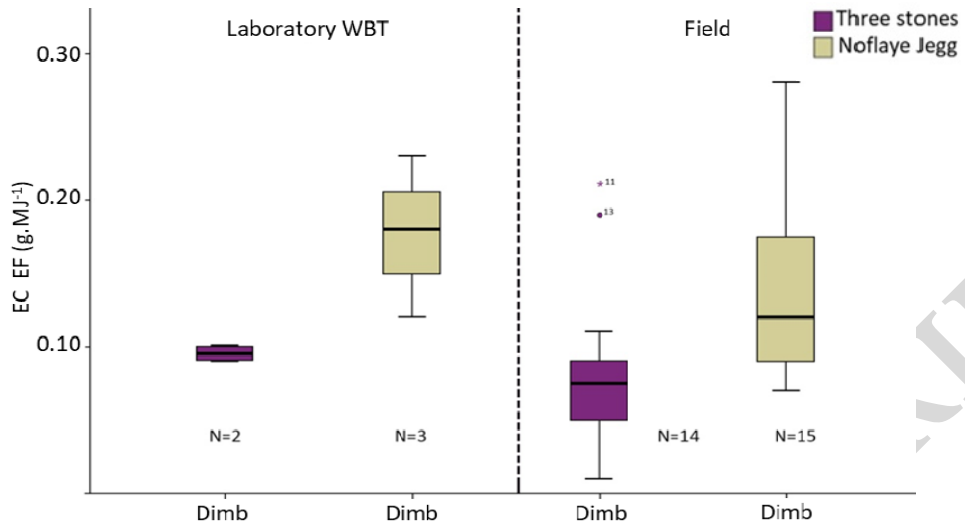
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727 Fig. 4.



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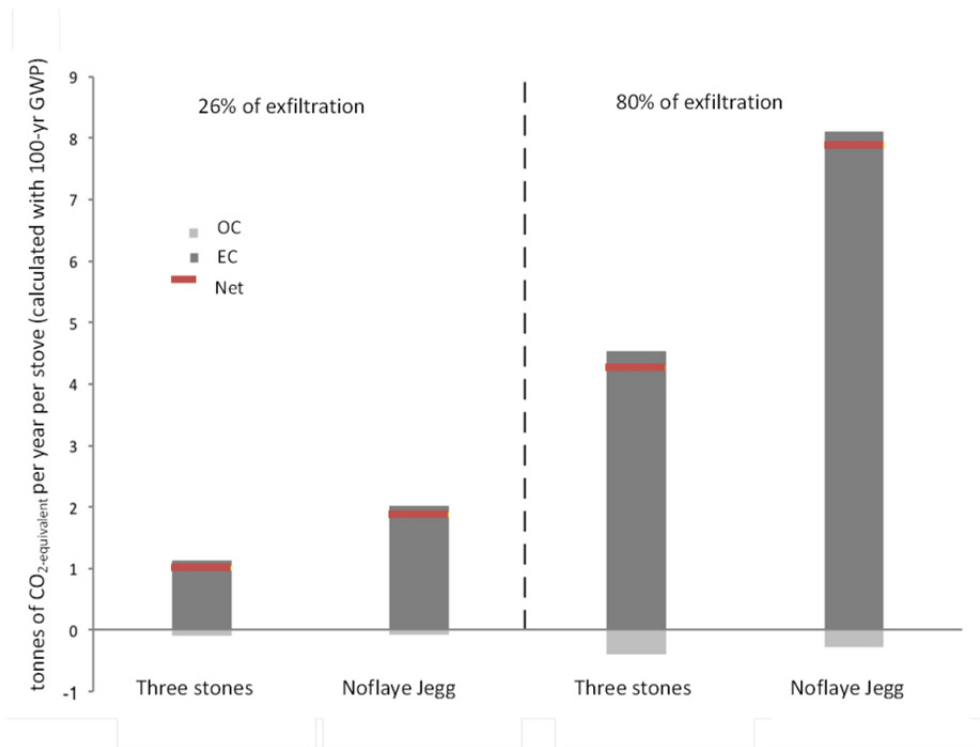
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742 Fig. 5



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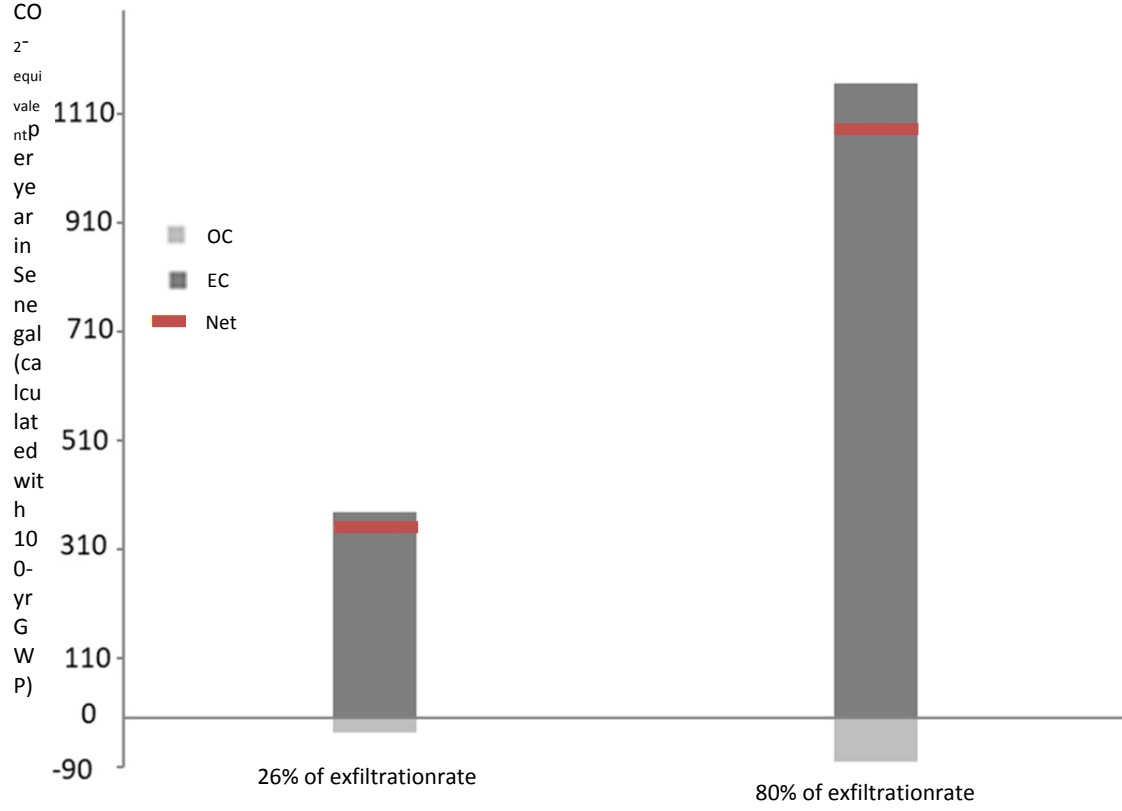
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756 Fig. 6

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