



## Reducing Air Pollutant Emissions from Burning Incense with the Addition of Calcium Carbonate

Chi-Ru Yang<sup>1\*</sup>, Ta-Chang Lin<sup>2</sup>, Yen-Shun Peng<sup>3</sup>, Sun-Zone Lee<sup>1</sup>, Yih-Feng Chang<sup>4</sup>

<sup>1</sup> Department of Environmental Engineering and Science, Chia-Nan University of Pharmacy and Science, Tainan 71710, Taiwan

<sup>2</sup> Department of Environmental Engineering National Cheng Kung University, Tainan 70101, Taiwan

<sup>3</sup> Department of Environmental Resources Management, Chia-Nan University of Pharmacy and Science, Tainan 71710, Taiwan

<sup>4</sup> Institute of Hot Spring Industry, Chia-Nan University of Pharmacy and Science, Tainan 71710, Taiwan

---

### ABSTRACT

A laboratory-scale study was performed to quantify the pollutant reduction effects from burning incense with the addition of CaCO<sub>3</sub>. Many studies have investigated the effects of burning incense on the quality of surrounding air, focusing primarily on particulate matter (PM) and polycyclic aromatic hydrocarbons (PAHs). However, the reduction of PM and PAHs from burning incense has received little attention. In our past study, we investigated nine types of commercially available incense and found that incense with a higher CaCO<sub>3</sub> content had lower PM and PAH emissions factors. Five to thirty percent of CaCO<sub>3</sub> was added to Liao and Chen incense powder, which are popular incense materials. The experimental results indicate that the reductions in the emissions of PM and PAHs from burning incense increased with along with amount of CaCO<sub>3</sub> additive. Mean PM reductions for 5.0%, 10.0%, 20.0%, and 30.0% CaCO<sub>3</sub> were 11 ± 2%, 15 ± 3%, 27 ± 1%, and 41 ± 3%, respectively. Mean particle-phase PAHs (P-PAHs) reductions were 9 ± 9%, 15 ± 5%, 22 ± 1%, and 28 ± 1%, respectively, and 5 ± 6%, 21 ± 1%, 21 ± 3%, and 30 ± 2% for total benzo[a]pyrene equivalent concentration (total BaP<sub>eq</sub>), respectively. This study was performed to quantify the reduction of PM and PAH emissions from burning incense with increasing amounts of CaCO<sub>3</sub>. The findings of this study may serve as a guide to producing safer and less-polluting incense.

**Keywords:** Incense; Calcium carbonate; Particulate matter; Polycyclic aromatic hydrocarbons; Total benzo[a]pyrene equivalent concentration.

---

### INTRODUCTION

It was estimated that there were 11,796 legal temples in Taiwan in 2009. The temple density was roughly 1 temple per 2,000 residents or 1 temple per 260 hectares; these figures exclude unregistered temples and altars (Department of Statistics, MOI, 2010). Incense is burned in temples and widely used during temple activities in Asia. There are 6.1 million families in Taiwan, with 45% burning incense twice per day (Lung *et al.*, 2007). Cancer, asthma, dermatitis, and genotoxic effects are related to exposure to incense smoke (Dawod and Hussain, 1995; Yang *et al.*, 1997; Jetter *et al.*, 2002). Navasumrit *et al.* (2008) indicated that exposure to carcinogens emitted from incense burning may increase health risk for the development of cancer in temple workers.

The burning of solid biomaterial such as incense, joss paper, tobacco, and wood has been demonstrated to significantly generate particulate matter (PM) (Fang *et al.*, 2003; Choosong *et al.*, 2010; Ning and Sioutas, 2010), polycyclic aromatic hydrocarbons (PAHs) (Yang *et al.*, 2007; Lo *et al.*, 2011), and polychlorinated dibenzo-*p*-dioxin/dibenzofurans (PCDD/Fs) (Wu *et al.*, 2010; Chiu *et al.*, 2011; Lin *et al.*, 2011).

The PM emission factor from burning incense ranges from 41 to 54 mg/g with an average of 46 mg/g, which is higher than that for cigarettes (10 mg/g) (Löfroth *et al.*, 1991; Mannix *et al.*, 1996). Burning incense in an enclosed room results in a suspended particle concentration of 390–730 µg/m<sup>3</sup>, which is 4–7 times higher than the indoor air particulate standard of the Taiwan EPA (100 µg/m<sup>3</sup>) (Kao and Lung, 2000). Cheng *et al.* (1995) found that the count median diameter (CMD) and mass median aerodynamic diameter (MMAD) of smoke aerosol from incense burning in an enclosed chamber were 0.13 and 0.28 µm, respectively. In our previous study, it was found that the distribution profiles of P-PAHs are similar to those of particle sizes (Yang *et al.*, 2007). Fine particles with diameters of less

---

\* Corresponding author. Tel.: +886 6 26649111 ext. 6337;  
Fax: +886 6 2669090  
E-mail address: cryang@mail.chna.edu.tw

than 2.5  $\mu\text{m}$  had higher specific surface areas and toxic effect than those of coarse particles with diameters of 2.5  $\mu\text{m}$  to 10  $\mu\text{m}$ . Moreover, small particles are more likely to harm the respiratory system as they can easily be inhaled and deposited in the respiratory tract and alveolar region (Harrison *et al.*, 2000; USEPA, 2002; Voutsas and Samara, 2002). The efficiency of particle deposition in the respiratory tract is a function of the particle size (Pope *et al.*, 1995; McAughey, 1997).

Several studies concluded that PAHs in indoor air mainly originated from incense burning in temples (Chiang and Liao, 2006; Lu *et al.*, 2008). In our previous study, gas-phase PAH (G-PAH) and P-PAH emission factors ranged from 10 to 29 and 4.5 to 6.9  $\mu\text{g/g}$ -incense, respectively. However, the particle-phase BaP<sub>eq</sub> emission factor was found to be consistently more than 40-fold higher than that of the corresponding gas-phase BaP<sub>eq</sub>. These results clearly suggest that in terms of carcinogenic potency, the control of P-PAH emissions is more important than the control of G-PAHs (Yang *et al.*, 2007). Many studies have suggested that PAHs are environmental immunosuppressive contaminants. PAHs, especially benzo(a)pyrene, not only harm the respiratory and immune system but also cause cell mutation and cancer, including lung and skin cancer (Hecht, 1999; Knize *et al.*, 1999; Laupeze *et al.*, 2002; Page *et al.*, 2002; Yousef *et al.*, 2002; van Grevenynghe *et al.*, 2003). Unfortunately, the reduction of PAHs and PM from burning incense has received little attention.

In our previous study, we investigated nine types of incense and found that incense with higher CaCO<sub>3</sub> content had lower PM and PAH emissions (Yang *et al.*, 2006). However, there were variations in the proportion of bamboo, adhesive and wood flour in incense sticks, the size of powder, and the composition of individual additives due to manufacturing differences. All of these factors influence the characteristics of air pollution from burning incense. Therefore, the incense used in the present study was made in our laboratory. In order to control the variations of incense characteristics, we added CaCO<sub>3</sub> and identified the reduction of PM and PAHs emissions in burning incense. The findings of this study are useful for safer and less-polluting incense produced.

## METHODS

### Manufacture of Test Incense

The main components of incense are powder and stick. Incense powders used in this study contains three ingredients including wood flour, adhesive (*Machilus kusanoi* Hay, a species of *Lauraceae*) and additives. Incense stick here is a bamboo stick (*Phyllostachys makinoi* Hay) which is a general

specification of incense manufacturers (length: 39.5 cm, weight: 0.55 g). Incense is named after its wooden materials; popular incense includes Liao and Chen. Liao incense is made from Chinese medical herbs. Chen incense was made from the several years' accumulation of natural resin which is caused by germs when the wood (*A. malaccensis*, a genus of *Aquilaria*) is hurt. The incense was partially handmade to reduce experimental error and to keep the process consistent. Each batch of incense used 100 g of powder. The proportion of adhesive in the powder was 20.0% (20.0 g). Besides, the percentages by weight of CaCO<sub>3</sub> in the powder were 5.0% (5.0 g), 10.0% (10.0 g), 20% (20.0 g), and 30% (30.0 g). The powder was then mixed with 100 g of deionized water. Aquiferous powder was pressed onto the bamboo stick with a hydraulic press machine to make semi-finished incense. The finished samples were conditioned in a carriage at 25°C under a relative humidity of 50% for 24 hrs before being weighed. The weights of the samples were 1.00  $\pm$  0.02 g. The schematic diagram of test incense stick is illustrated in Fig. 1. Each incense stick base part was 11.5 cm, 0.16 g of bamboo, burned part was 28 cm, 0.84 g, including 0.45 g of powder (adhesive 0.09 g) and 0.39 g of bamboo. The detailed compositions of the test incense with various amounts of CaCO<sub>3</sub> additive are listed in Table 1.

### Sampling Program

The sampling set-up is illustrated in Fig. 2 (Yang *et al.*, 2006). The air exchange rate (Ach) in a domestic environment in Taiwan ranges from 0.8 to 3.5 Ach in summer and 0.5 to 2.0 Ach in winter (Li and Ro, 2000). In this study, the air exchange rate was maintained at 1.5 Ach to simulate natural adequate ventilation conditions. All experiments were conducted in a 1.2-m<sup>3</sup> stainless steel #304 environmental test chamber with a 30 L air/min flow rate. An air-cleaning train, consisting of a high-efficiency particulate air (HEPA) filter set, followed immediately by an activated carbon bed and a XAD-2 resin bed, was used to provide clean air. The train sequentially removed particulates, gaseous components, and organic substances in the feed air before it was drawn into the combustion chamber. Before each run, the chamber was first flushed thoroughly with clean air provided by the above-mentioned system. In order to prevent the air from directly blowing on the burning incense, the air was forced to pass through packing rings before entering the combustion room. Finally, the outlet air was pumped out of the laboratory. A panel-mounted flow meter was installed in front of a 187-W (1/4-hp) air pump. To avoid possible errors in flow measurement, the panel-mounted flow meter was recalibrated with an infrared ray soap bubble calibrator (Gilibrator-2, Gilian Instrument Corp.) after every 5–7 runs. A small low-speed fan was placed inside the chamber, which blew towards

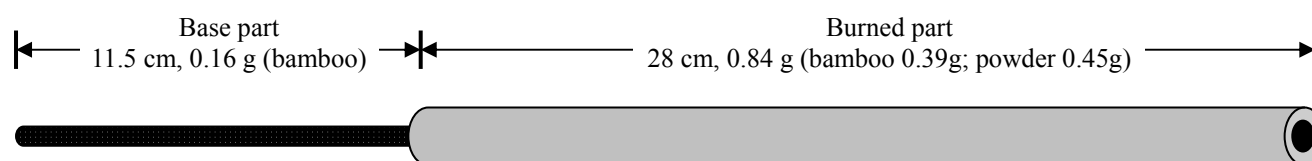


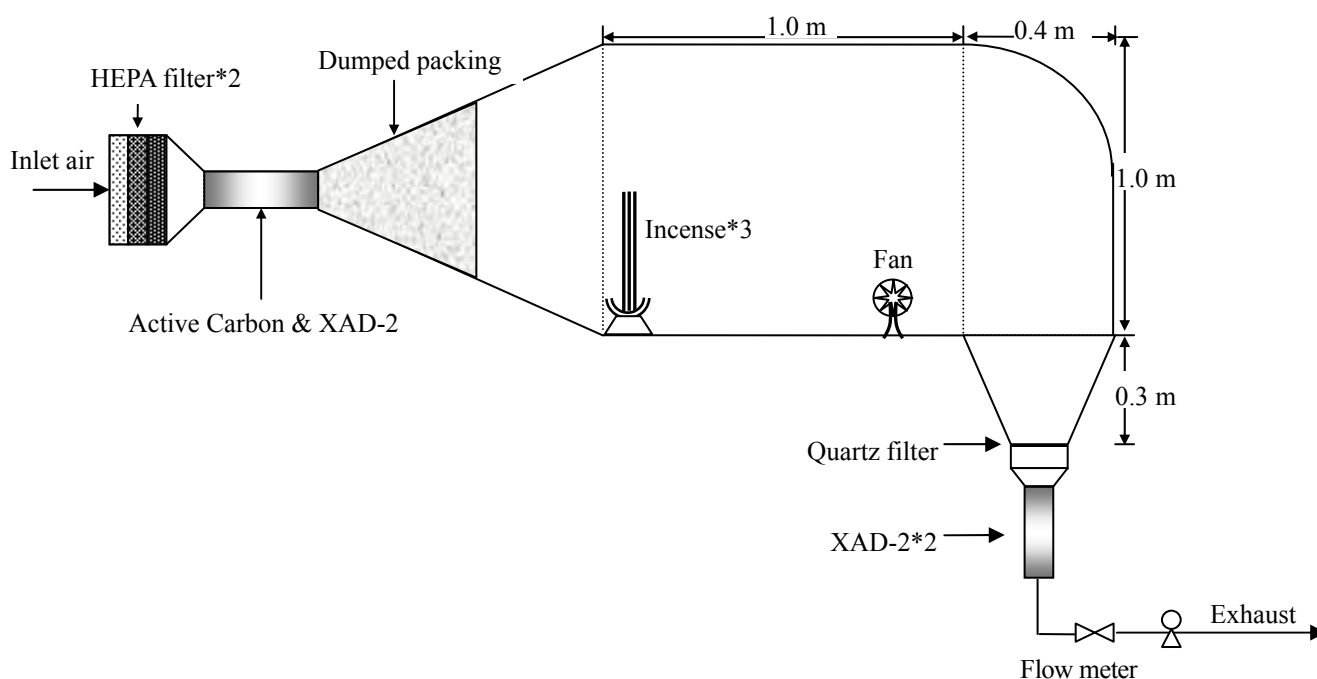
Fig. 1. Schematic diagram of test incense stick.

**Table 1.** Composition of test incense stick for various amounts of CaCO<sub>3</sub> additive.

Test incense <sup>a</sup>	CaCO <sub>3</sub> in powder (%)	Powder (g) = 0.45 g			Bamboo (g) <sup>b</sup>
		CaCO <sub>3</sub>	Wood flour	Adhesive	
CC00	0	0	0.36	0.09	0.55
CC05	5.0	0.0225	0.4275	0.09	0.55
CC10	10.0	0.045	0.405	0.09	0.55
CC20	20.0	0.09	0.36	0.09	0.55
CC30	30.0	0.135	0.315	0.09	0.55

a: the weight of one stick of all types of test incense was 1.00 g (0.45 g powder and 0.55 g bamboo stick).

b: the weight of each stick bamboo burned was 0.39 g.

**Fig. 2.** Schematic diagram of the burning chamber with sampling attachments.

a corner of the ceiling of the chamber (away from the incense) to stir the air without directly affecting the rising plume of smoke. This study monitored the temperature maintained below 35°C during sampling in the chamber. We set up replicated tests for 10 types of incense and each of them was performed six times.

Incense sticks (three sticks per run) were placed upright on a pre-weighed metallic plate. A ring stand held the samples in the rear of the chamber. At the beginning of each run, the incense sticks were positioned so that their tips were located roughly at the half-height of the chamber. The incense sticks were briefly lit with a propane lighter. The flame was then immediately put out to initiate smoldering. To ensure quantitative collection of all smoke generated during each run, clean air (pumped at a flow rate four times that used for incense burning) was continuously drawn to purge the chamber for four additional hours after the combustion was completed.

The quartz filters (Gelman Series, 100 mm in diameter) were pretreated before sampling by heating in a muffle furnace in air for 2.5 h at 900°C. Each quartz filter was cleaned by extraction with a mixed solvent (1:1 n-hexane and

dichloromethane) for 24 hrs in a Soxhlet extractor. The filters were conditioned for 24 hrs at constant 25°C and a relative humidity of 50% in a dry box before and after sampling.

#### Analyses of PAHs

After final weighing, all filters were separately placed in appropriate Soxhlet extractors and extracted with 600 mL of a dichloromethane /n-hexane mixture (v/v = 1:1) for 24 hrs. The extract was then concentrated under ultra-pure nitrogen, cleaned, and re-concentrated to exactly 1.0 mL. All extracts were analyzed with a gas chromatograph/mass selective detector (GC/MSD) (GC-6890N with MSD-5973, Agilent Technologies, USA) with a J&W Ultra2 capillary column (50 m × 0.314 mm × 0.17 μm). A computer-controlled automatic sampler (Model 3365, Hewlett Packard, USA) was used in conjunction with the GC/MSD system. All injections were splitless with an injection volume of 1 μL. The injector and the detector temperatures were 300 and 325°C, respectively. The temperature program included an immediate fast initial increase from 50 to 100°C at 20 °C/min, followed by a milder increase from 100 to 290°C at 3 °C/min, and finally a hold at 290°C for 20 minutes.

The concentrations of the following PAHs were determined: naphthalene (Nap) for 2-ring; acenaphthylene (AcPy), acenaphthene (Acp), fluorene (Flu), phenanthrene (PA), and anthracene (Ant) for 3-ring; fluoranthene (FL), pyrene (Pyr), benzo[*a*]anthracene (BaA), and chrysene (CHR) for 4-ring; benzo[*b*]fluoranthene (BbF), benzo[*k*]fluoranthene (BkF), benzo[*a*]pyrene (BaP), dibenzo[*a,h*]anthracene (DBA) for 5-ring; and indeno[1,2,3-*cd*]pyrene (IND), benzo[*ghi*]perylene (Bghip) for 6-ring.

The GC/MSD was calibrated with a diluted standard solution of 16 PAH compounds (PAHs mixture-610M from Supelco). The standard solution concentrations were 0.001, 0.005, 0.01, 0.05, 0.1, 0.5, 1, 5 and 10  $\mu\text{g/mL}$ . PAHs recovery efficiencies were determined by processing a solution containing known PAHs concentrations through the same experimental procedure used for the samples. The recovery efficiencies of PAHs varied from 69.1 (Nap) to 98.3 (BaA), with an average of 87.6%. The method detection limit (MDL) for the 16 PAHs, including Nap, Acpy, Acp, Flu, Pa, Ant, Fl, Pyr, BaA, CHR, BbF, BkF, BaP, IND, DBA and Bghip, were found to be 0.560, 0.531, 0.430, 0.623, 0.832, 0.071, 0.319, 0.098, 0.244, 0.542, 0.405, 0.274, 0.542, 0.273, 0.936, and 0.443 ppb, respectively. Ten consecutive injections of a PAHs 610-M standard yielded an average relative standard deviation (RSD) of the GC integration area of 3.0%, with a range of 0.8–5.1%. The blank tests for PAHs were accomplished using the same procedure as that used for the recovery-efficiency tests without adding the known standard solution before extraction. Analyses of field blanks, including the blank quartz filter and glass bottle showed no significant contamination (GC/MSD integrated area < detection limit). The calculation procedures of the incense PAH results were not modified with the recovery efficiencies and the field blanks. The recovery efficiencies of the 16 individual PAH compounds and the field blanks were not used to modified the PAH results.

#### Data Analysis

The 16 individual PAHs were divided according to their molecular weight into three categories: low molecular weight (LM-PAHs containing two- and three-ringed PAHs); middle molecular weight (MM-PAHs containing four-ringed PAHs); and high molecular weight (HM-PAHs containing five- and six-ringed PAHs). The total PAHs concentration was the sum of the concentrations for the 16 PAH compounds in each collected sample. Moreover, considering that several PAH compounds are known human carcinogens, the carcinogenic potencies of PAHs emissions from each emission source were

also determined. In principle, the carcinogenic potency of a given PAH compound is assessed on the basis of its  $\text{BaP}_{\text{eq}}$ . The calculation of the  $\text{BaP}_{\text{eq}}$  concentration for a given PAH compound is determined by its toxic equivalent factor (TEF), which represents the relative carcinogenic potency of the given PAH compound, using benzo[*a*]pyrene as a reference compound to adjust its original concentration. This study applied the TEFs completed by Nisbet and LaGoy (1992) to assess the carcinogenic potency of total PAHs (i.e., total  $\text{BaP}_{\text{eq}}$ ) using the sum of the  $\text{BaP}_{\text{eq}}$  concentrations estimated for each PAH compound with a TEF in the total PAHs.

## RESULTS AND DISCUSSION

### Burning Time and Burning Rate of Test Incense

The effect of  $\text{CaCO}_3$  additive on the burning time and burning rate of the test incense, Liao incense and Chen incense, is shown in Table 2. For Liao incense, the mean reductions of burning time for CC05, CC10, CC20, and CC30  $\text{CaCO}_3$  additives were  $4.3 \pm 3.2\%$ ,  $8.5 \pm 3.2\%$ ,  $22 \pm 3.2\%$ , and  $30 \pm 3.2\%$ , respectively. For Chen incense, the mean reductions were  $5.7 \pm 5.7\%$ ,  $9.2 \pm 6.9\%$ ,  $24 \pm 3.4\%$ , and  $39 \pm 3.4\%$ , respectively. The mean reductions of burning time were  $5.0 \pm 1.1\%$ ,  $8.9 \pm 0.5\%$ ,  $23 \pm 1.3\%$ , and  $34 \pm 6.6\%$ , respectively, for CC05, CC10, CC20, and CC30. In this study, we controlled the process variations and added accurate amounts of  $\text{CaCO}_3$  to help verify the reduction of burning time. The mean burning time of the two types for CC30 incense was less than 60 minutes. Users prefer incense with a long burning time. The appropriate burning time was important. The results can be applied to the commercial production of incense.

These results may be attributed to  $\text{CaCO}_3$ , which has a high boiling point;  $\text{CaCO}_3$  may trap the heat energy generated at the burning tip during combustion. Moreover,  $\text{CaCO}_3$  with refractory characteristics in terms of low heat transfer coefficient (0.05 W/mol·K) and high molar heat capacity (81.88 J/mol·K) will prevent air convection and thus raise the burning rate of inflammable substances. Compared with the result conducted by Yang (Yang et al., 2006), the reasons for high burning rate observed in this study included different raw materials, smaller diameter incense (2.2 mm in this study, 2.5–3.0 mm in previous study), incense density changes resulted from incense production (semi-handmade in this study, machine-made in previous study).

### PM Emission Factor

The PM emission factor results are shown in Fig. 3. It was

**Table 2.** Burning time and burning rate of test incense for various amounts of  $\text{CaCO}_3$  additive.

Test incense	$\text{CaCO}_3$ in burned part (%)	Burning time (min)		Burning rate (mg/min)	
		Liao	Chen	Liao	Chen
CC00	0	$94 \pm 1$	$87 \pm 2$	$27 \pm 1$	$29 \pm 1$
CC05	2.68	$90 \pm 2$	$82 \pm 3$	$27 \pm 1$	$31 \pm 1$
CC10	5.36	$86 \pm 2$	$79 \pm 4$	$29 \pm 0$	$32 \pm 2$
CC20	10.7	$73 \pm 2$	$66 \pm 1$	$35 \pm 1$	$38 \pm 0$
CC30	16.1	$66 \pm 2$	$53 \pm 1$	$38 \pm 1$	$47 \pm 1$

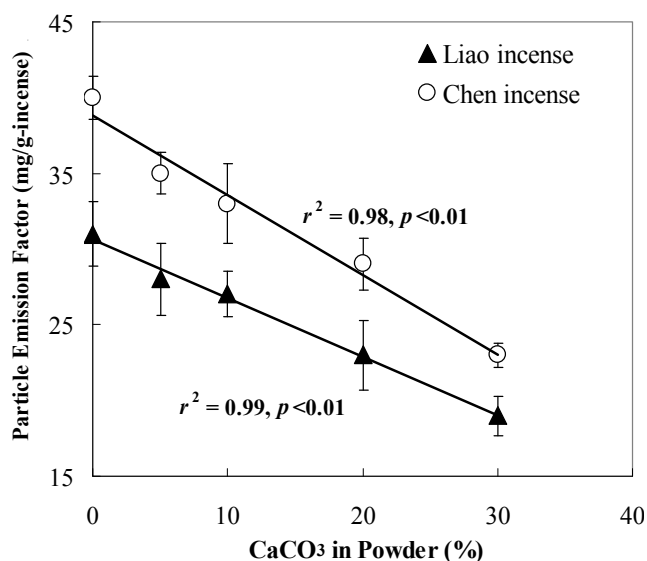
N = 6 for each type of incense.

found that the PM emission factor values of original (CC00) Liao incense and Chen incense were  $31 \pm 2.1$  and  $40 \pm 1.4$  mg/g-incense, respectively. For Liao incense, the mean PM reductions for CC05, CC10, CC20, and CC30 were  $10 \pm 14\%$ ,  $13 \pm 12\%$ ,  $26 \pm 14\%$ , and  $39 \pm 11\%$ , respectively. For Chen incense, the mean reductions were  $13 \pm 7.0\%$ ,  $18 \pm 10\%$ ,  $28 \pm 8.3\%$ , and  $43 \pm 5.5\%$ , respectively. These results indicate that the reduction of PM emissions increases with increasing amount of  $\text{CaCO}_3$  additive.

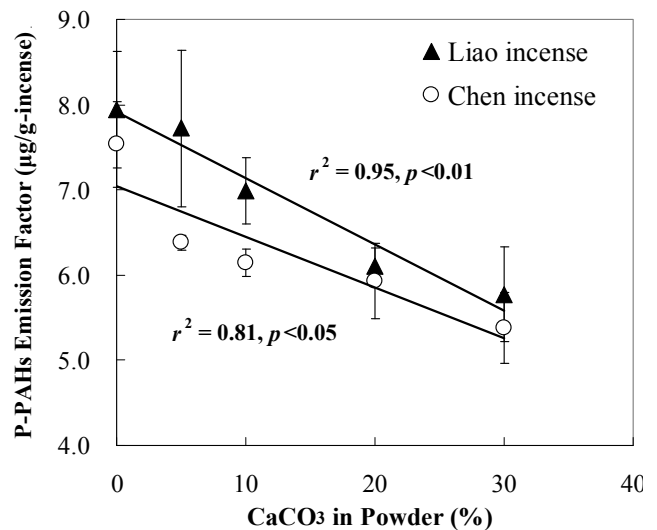
The correlations of the PM emission factor and the  $\text{CaCO}_3$  in powder with regression analysis ( $r^2 = 0.99$ ,  $p < 0.01$  for Liao;  $r^2 = 0.98$ ,  $p < 0.01$  for Chen) were strongly negative. Thus, the more addition of  $\text{CaCO}_3$  in powder will emitted less the particulate emissions in burning incense. Moreover, the mean PM reductions for the two types of test incense were  $11 \pm 2\%$ ,  $15 \pm 3\%$ ,  $27 \pm 1\%$ , and  $41 \pm 3\%$ , respectively, for CC05, CC10, CC20, and CC30, which are higher than the  $\text{CaCO}_3$  content was 2.68%, 5.36%, 10.7%, and 16.1% in the burned part, respectively. These results indicate that the addition of  $\text{CaCO}_3$  significantly reduces PM emissions. This may be attributed to  $\text{CaCO}_3$ , which acted as noncombustible replacer to reduce emission from organic wood material, such as bamboo, adhesive, and wood flour in the burned incense.

#### PAH Emission Factor

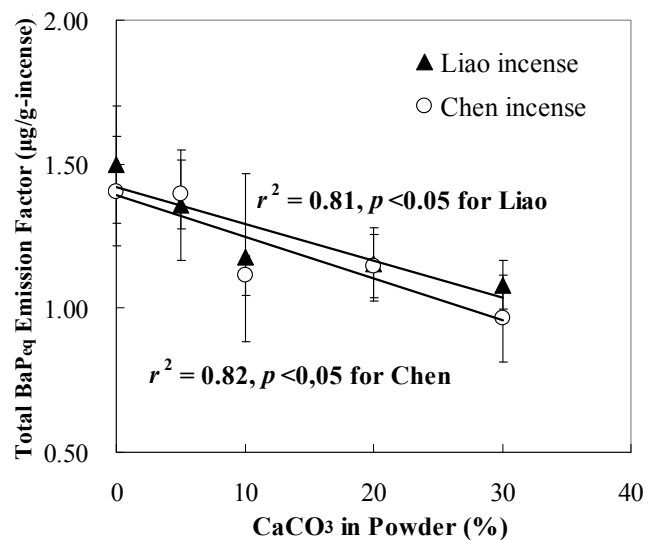
The emission factor results of P-PAHs are shown in Fig. 4. It was found that the P-PAHs emission factors for original Liao incense and Chen incense were  $7.9 \pm 0.7$  and  $7.5 \pm 0.5$   $\mu\text{g/g-incense}$ , respectively. For Liao incense, the mean P-PAHs reductions for CC05, CC10, CC20, and CC30 were  $3 \pm 20\%$ ,  $12 \pm 14\%$ ,  $23 \pm 11\%$ , and  $27 \pm 16\%$ , respectively. For Chen incense, the mean reductions were  $15 \pm 8.0\%$ ,  $18 \pm 9.1\%$ ,  $21 \pm 13\%$ , and  $29 \pm 12\%$ , respectively. Similar results were found for total  $\text{BaP}_{\text{eq}}$ . The emission factor results of total  $\text{BaP}_{\text{eq}}$  are shown in Fig. 5. The total  $\text{BaP}_{\text{eq}}$  emission factor values for original Liao incense and Chen incense



**Fig. 3.** PM emission factor of test incense for various amounts of  $\text{CaCO}_3$  additive (Each error value equals one standard deviation)



**Fig. 4.** P-PAH emission factor of test incense for various amounts of  $\text{CaCO}_3$  additive (Each error value equals one standard deviation).



**Fig. 5.** Total  $\text{BaP}_{\text{eq}}$  emission factor of test incense for various amounts of  $\text{CaCO}_3$  additive (Each error value equals one standard deviation).

were  $1.5 \pm 0.2$  and  $1.4 \pm 0.2$   $\mu\text{g/g-incense}$ , respectively. For Liao incense, the mean total  $\text{BaP}_{\text{eq}}$  reductions for CC05, CC10, CC20, and CC30 were  $9 \pm 22\%$ ,  $22 \pm 31\%$ ,  $23 \pm 19\%$ , and  $28 \pm 17\%$ , respectively. For Chen incense, the mean reductions were  $1 \pm 22\%$ ,  $21 \pm 19\%$ ,  $18 \pm 21\%$ , and  $31 \pm 24\%$ , respectively. These results indicate that P-PAHs and total  $\text{BaP}_{\text{eq}}$  emissions from incense can be reduced using  $\text{CaCO}_3$  additive; the reduction increased with increasing amount of  $\text{CaCO}_3$  additive.

The correlations of the P-PAHs emission factor and the  $\text{CaCO}_3$  in powder with regression analysis ( $r^2 = 0.95$ ,  $p < 0.01$  for Liao;  $r^2 = 0.81$ ,  $p < 0.05$  for Chen) were strongly negative. In addition, the correlations of the total  $\text{BaP}_{\text{eq}}$  emission factor and the  $\text{CaCO}_3$  in powder with regression analysis ( $r^2 = 0.81$ ,  $p < 0.05$  for Liao;  $r^2 = 0.82$ ,  $p < 0.05$  for

Chen) were strongly negative. Thus, the more addition of CaCO<sub>3</sub> in powder will emitted less the P-PAHs and total BaP<sub>eq</sub> emissions in burning incense.

As mentioned, CaCO<sub>3</sub> content was 2.68%, 5.36%, 10.7%, and 16.1% in the burned part, respectively, for CC05, CC10, CC20, and CC30, which are less than the means of P-PAHs reductions ( $9 \pm 9\%$ ,  $15 \pm 5\%$ ,  $22 \pm 1\%$ , and  $28 \pm 1\%$ , respectively) and total BaP<sub>eq</sub> reductions ( $5 \pm 6\%$ ,  $21 \pm 1\%$ ,  $21 \pm 3\%$ , and  $30 \pm 2\%$ , respectively) for the two types of test incense. These results strongly suggest that the addition

of CaCO<sub>3</sub> significantly reduces P-PAHs emissions. These results may be attributed to CaCO<sub>3</sub> decreased the smolder effect during incense combustion. Therefore, the addition of CaCO<sub>3</sub> efficiently decreases genotoxic P-PAHs.

### Profiles of PAHs

The 16 individual PAHs were divided according to their molecular weight into three categories: LM-PAHs, MM-PAHs and, HM-PAHs. Table 3 and Fig. 6 show that the emission factor values for Liao incense were  $0.97 \pm 0.31$ ,

**Table 3.** Individual 16 P-PAH emission factors of test incense for various amounts of CaCO<sub>3</sub> additive.

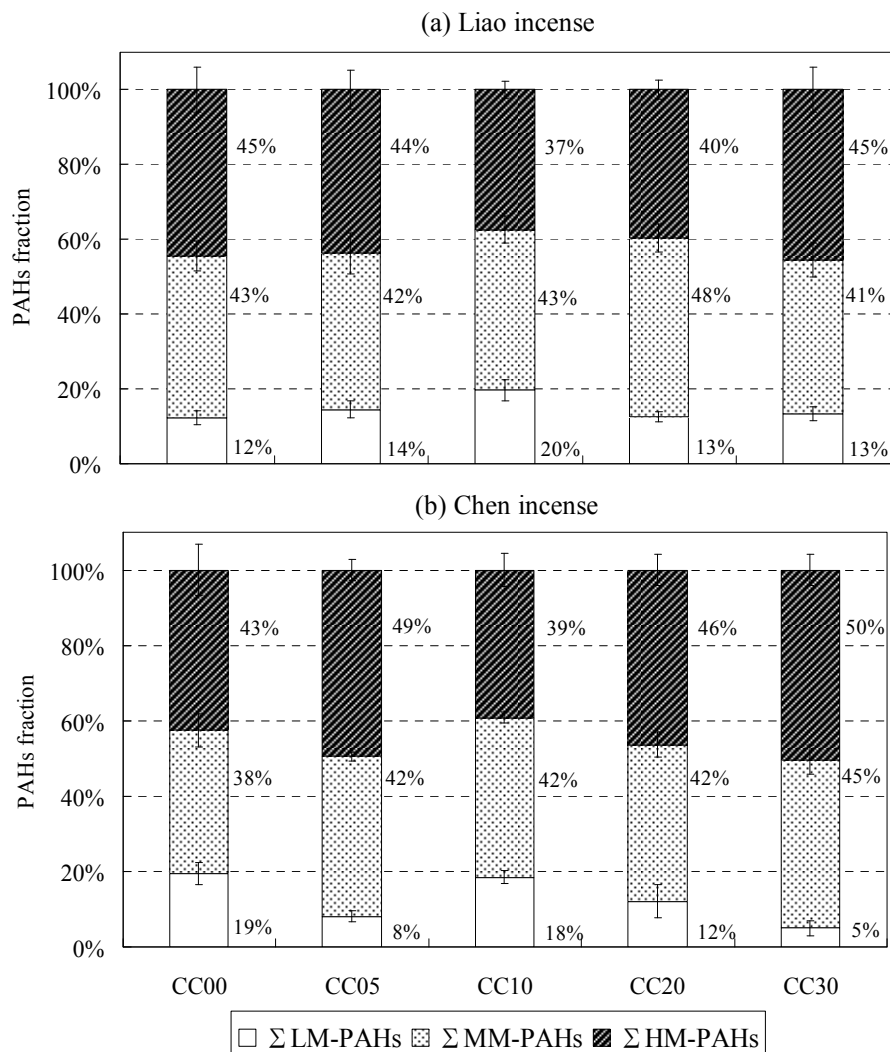
P-PAHs ( $\mu\text{g/g-incense}$ )	Liao incense					Chen incense				
	CC00	CC05	CC10	CC20	CC30	CC00	CC05	CC10	CC20	CC30
Nap	0.11 $\pm 0.04$	0.09 $\pm 0.06$	0.09 $\pm 0.06$	0.05 $\pm 0.04$	0.05 $\pm 0.03$	0.15 $\pm 0.04$	0.02 $\pm 0.07$	0.09 $\pm 0.05$	0.08 $\pm 0.03$	0.03 $\pm 0.04$
Acpy	0.12 $\pm 0.06$	0.11 $\pm 0.04$	0.27 $\pm 0.10$	0.10 $\pm 0.04$	0.09 $\pm 0.03$	0.30 $\pm 0.08$	0.03 $\pm 0.23$	0.21 $\pm 0.11$	0.08 $\pm 0.05$	0.02 $\pm 0.04$
Acp	0.02 $\pm 0.02$	0.02 $\pm 0.01$	0.03 $\pm 0.01$	0.01 $\pm 0.03$	0.01 $\pm 0.02$	0.03 $\pm 0.00$	0.01 $\pm 0.01$	0.03 $\pm 0.00$	0.02 $\pm 0.03$	0.01 $\pm 0.01$
Flu	0.04 $\pm 0.01$	0.03 $\pm 0.05$	0.04 $\pm 0.03$	0.03 $\pm 0.04$	0.03 $\pm 0.05$	0.04 $\pm 0.02$	0.02 $\pm 0.02$	0.04 $\pm 0.01$	0.03 $\pm 0.01$	0.01 $\pm 0.03$
Pa	0.49 $\pm 0.21$	0.55 $\pm 0.32$	0.68 $\pm 0.33$	0.42 $\pm 0.26$	0.39 $\pm 0.37$	0.65 $\pm 0.17$	0.32 $\pm 0.22$	0.56 $\pm 0.25$	0.36 $\pm 0.21$	0.16 $\pm 0.17$
Ant	0.19 $\pm 0.03$	0.32 $\pm 0.05$	0.26 $\pm 0.02$	0.16 $\pm 0.06$	0.20 $\pm 0.05$	0.29 $\pm 0.01$	0.13 $\pm 0.03$	0.20 $\pm 0.02$	0.15 $\pm 0.05$	0.04 $\pm 0.02$
Fl	0.63 $\pm 0.26$	0.64 $\pm 0.34$	0.64 $\pm 0.22$	0.60 $\pm 0.32$	0.47 $\pm 0.13$	0.61 $\pm 0.19$	0.49 $\pm 0.23$	0.51 $\pm 0.18$	0.46 $\pm 0.21$	0.41 $\pm 0.18$
Pyr	0.67 $\pm 0.10$	0.68 $\pm 0.08$	0.62 $\pm 0.10$	0.62 $\pm 0.14$	0.50 $\pm 0.09$	0.61 $\pm 0.08$	0.53 $\pm 0.03$	0.50 $\pm 0.07$	0.47 $\pm 0.08$	0.42 $\pm 0.06$
BaA	0.81 $\pm 0.31$	0.73 $\pm 0.18$	0.64 $\pm 0.21$	0.82 $\pm 0.37$	0.53 $\pm 0.19$	0.65 $\pm 0.32$	0.65 $\pm 0.29$	0.59 $\pm 0.16$	0.57 $\pm 0.26$	0.57 $\pm 0.22$
CHR	1.33 $\pm 0.47$	1.17 $\pm 0.52$	1.10 $\pm 0.34$	0.86 $\pm 0.35$	0.86 $\pm 0.31$	1.00 $\pm 0.41$	1.04 $\pm 0.41$	0.99 $\pm 0.42$	0.98 $\pm 0.46$	1.01 $\pm 0.37$
BbF	0.54 $\pm 0.16$	0.48 $\pm 0.12$	0.46 $\pm 0.13$	0.41 $\pm 0.11$	0.37 $\pm 0.08$	0.52 $\pm 0.07$	0.54 $\pm 0.17$	0.40 $\pm 0.12$	0.47 $\pm 0.12$	0.51 $\pm 0.10$
BkF	0.34 $\pm 0.08$	0.43 $\pm 0.06$	0.22 $\pm 0.03$	0.22 $\pm 0.02$	0.26 $\pm 0.04$	0.39 $\pm 0.04$	0.27 $\pm 0.08$	0.19 $\pm 0.03$	0.28 $\pm 0.02$	0.28 $\pm 0.03$
BaP	1.11 $\pm 0.25$	1.00 $\pm 0.22$	0.82 $\pm 0.17$	0.84 $\pm 0.31$	0.80 $\pm 0.22$	0.97 $\pm 0.38$	1.05 $\pm 0.43$	0.81 $\pm 0.33$	0.84 $\pm 0.27$	0.64 $\pm 0.38$
IND	0.71 $\pm 0.15$	0.67 $\pm 0.14$	0.51 $\pm 0.06$	0.50 $\pm 0.06$	0.55 $\pm 0.11$	0.59 $\pm 0.12$	0.61 $\pm 0.06$	0.49 $\pm 0.13$	0.57 $\pm 0.06$	0.62 $\pm 0.12$
DBA	0.13 $\pm 0.06$	0.10 $\pm 0.08$	0.15 $\pm 0.09$	0.10 $\pm 0.07$	0.09 $\pm 0.07$	0.19 $\pm 0.07$	0.12 $\pm 0.10$	0.12 $\pm 0.06$	0.10 $\pm 0.09$	0.11 $\pm 0.07$
BghiP	0.70 $\pm 0.18$	0.69 $\pm 0.15$	0.44 $\pm 0.09$	0.35 $\pm 0.11$	0.56 $\pm 0.18$	0.53 $\pm 0.18$	0.58 $\pm 0.14$	0.40 $\pm 0.20$	0.50 $\pm 0.14$	0.55 $\pm 0.13$
$\Sigma\text{LM-PAHs}$	0.97 $\pm 0.31$	1.12 $\pm 0.28$	1.37 $\pm 0.42$	0.77 $\pm 0.19$	0.77 $\pm 0.21$	1.46 $\pm 0.17$	0.53 $\pm 0.24$	1.13 $\pm 0.22$	0.71 $\pm 0.28$	0.27 $\pm 0.12$
$\Sigma\text{MM-PAHs}$	3.44 $\pm 0.52$	3.22 $\pm 0.54$	3.00 $\pm 0.67$	2.90 $\pm 0.46$	2.36 $\pm 0.38$	2.87 $\pm 0.66$	2.71 $\pm 0.49$	2.59 $\pm 0.31$	2.47 $\pm 0.41$	2.41 $\pm 0.34$
$\Sigma\text{HM-PAHs}$	3.53 $\pm 0.47$	3.37 $\pm 0.41$	2.60 $\pm 0.18$	2.42 $\pm 0.18$	2.63 $\pm 0.36$	3.19 $\pm 0.39$	3.17 $\pm 0.21$	2.41 $\pm 0.33$	2.75 $\pm 0.27$	2.71 $\pm 0.25$
<b>Total PAHs</b>	7.94 $\pm 0.68$	7.72 $\pm 0.92$	6.99 $\pm 0.39$	6.10 $\pm 0.22$	5.78 $\pm 0.55$	7.53 $\pm 0.51$	6.38 $\pm 0.09$	6.14 $\pm 0.17$	5.93 $\pm 0.44$	5.38 $\pm 0.41$

Note: P-PAHs: Particle-phase PAHs.

LM-PAHs: Low Molecular Weight PAHs (2-3ring).

MM-PAHs: Median Molecular Weight PAHs (4rings).

HM-PAHs: High Molecular Weight PAHs (5-6rings).



**Fig. 6.** Fractions of P-PAH classifications of test incense for various amounts of CaCO<sub>3</sub> additive: (a) Liao incense; (b) Chen incense.

$3.44 \pm 0.52$ , and  $3.53 \pm 0.47$   $\mu\text{g/g-incense}$ , respectively, accounting for  $12 \pm 3.9\%$ ,  $43 \pm 6.6\%$ , and  $45 \pm 5.9\%$  of total P-PAHs. LM-PAH, MM-PAH, and HM-PAH emission factor values from Liao incense for CC05, CC10, CC20, and CC30 were 0.77–1.37, 2.36–3.22, and 2.42–3.37  $\mu\text{g/g-incense}$ , respectively, accounting for 13–20%, 41–48%, and 37–45% of total P-PAHs. For Chen incense, the values were 1.46, 2.87, and 3.19  $\mu\text{g/g-incense}$ , respectively, accounting for  $19 \pm 2.9\%$ ,  $38 \pm 11\%$ , and  $43 \pm 6.8\%$  of total P-PAHs. For Chen incense, the values were 0.27–1.13, 2.41–2.71, and 2.41–3.17  $\mu\text{g/g-incense}$ , respectively, accounting for 5.0–18%, 42–45%, and 39–50% of total P-PAHs.

In addition, according to Chi-Square test of homogeneity, the percentages of the rings of PAHs were quite similar to those of ten types of incense ( $p > 0.5$ ). On the other words, the PAHs emission profiles of the two original types of incense are similar. In addition, it may be concluded that individual PAH emission profiles of various types of incense are similar and thus independent of CaCO<sub>3</sub> content. The reason might be that although the incense with CaCO<sub>3</sub> burned less organic material, it was still a wood combustion type.

## CONCLUSIONS

Experimental results indicate that the addition of 10% CaCO<sub>3</sub> in powder reduced the burning time by  $10 \pm 1.2$  min, increased the burning rate by  $5.0 \pm 1.2$  mg/min, reduced PM by  $4.56 \pm 0.99$  mg/g-incense, and reduced P-PAHs by  $0.75 \pm 0.08$   $\mu\text{g/g-incense}$  and particle phase BaP<sub>eq</sub> by  $0.15 \pm 0.01$   $\mu\text{g/g-incense}$  with linear regression analysis. Summary of above-mentioned, these results may be attributed to a lower amount of organic wood materials, such as bamboo, adhesive, and wood flour, being burned in incense with CaCO<sub>3</sub> additives. CaCO<sub>3</sub> prevented air convection and kept a high temperature at the burning tip, which decreased the smolder effect during incense combustion. In addition, individual PAH emission profiles of various types of incense are similar and thus independent of CaCO<sub>3</sub> content. However, the quality of incense, in terms of fragrance emission rate and burning time, may be slightly compromised due to enhanced burning efficiency, the addition of CaCO<sub>3</sub> effectively reduces emissions that are harmful to human health. The findings of this study may serve as a guide for producing safer incense.

## ACKNOWLEDGEMENTS

The authors would like to thank the National Science Council of Taiwan for financially supporting this research under Grant No. NSC 96-2221-E-041-006.

## REFERENCES

- Cheng, Y.S., Bechtold, W.E., Yu, C.C. and Hung, I.F. (1995). Incense Smoke: Characterization and Dynamics in Indoor Environments. *Am. Ind. Hyg. Assoc. J.* 23: 271–281.
- Chiang, K.C. and Liao, C.M. (2006). Heavy Incense Burning in Temples Promotes Exposure Risk from Airborne PMs and Carcinogenic PAHs. *Sci. Total Environ.* 372: 64–75.
- Chiu, J.C., Shen, Y.H., Li, H.W., Lin, L.F., Wang, L.C. and Chang-Chien, G.P. (2011). Emissions of Polychlorinated Dibenzo-*p*-dioxins and Dibenzofurans from an Electric Arc Furnace, Secondary Aluminum Smelter, Crematory and Joss Paper Incinerators. *Aerosol Air Qual. Res.* 11: 13–20.
- Choosong, T., Hata, M. and Furuuchi, M. (2010). Workplace Environment and Personal Exposure of PM and PAHs to Workers in Natural Rubber Sheet Factories Contaminated by Wood Burning Smoke. *Aerosol Air Qual. Res.* 10: 8–21.
- Dawod, S.T. and Hussain, A.A. (1995). Childhood Asthma in Qatar. *Ann. Allergy Asthma Immunol.* 75: 360–364.
- Department of Statistics, Ministry of Interior (2010). General Conditions of Religions, Available at: <http://sowf.moi.gov.tw/stat/year/y01-03.xls>.
- Fang, G.C., Chang, C.N., Chu, C.C., Wu, Y.S., Fu, P.P.C., Chang, S.C. and Yang, I.L. (2003). Fine (PM<sub>2.5</sub>), Coarse (PM<sub>2.5-10</sub>), and Metallic Elements of Suspended Particulates for Incense Burning at Tzu Yun Yen Temple in Central Taiwan. *Chemosphere* 51: 983–991.
- Harrison, R.M., Shi, J.P., Xi, S., Khan, A., Mark, D., Kinnersley, R., and Yin, J. (2000). Measurement of Number, Mass and Size Distribution of Particles in the Atmosphere. *Philos. Trans. R. Soc. London, Ser. A* 358: 2567–2580.
- Hecht, S.S. (1999). Tobacco Smoke Carcinogens and Lung Cancer. *J. Nat. Cancer Inst.* 91: 1194–1210.
- Jetter, J.J., Guo, Z., McBrien, J.A. and Flynn, M.R. (2003). Characterization of Emissions from Burning Incense. *Sci. Total Environ.* 295: 51–67.
- Kao, M.C. and Lung, S.C. (2000). Personal Particulate Exposures in Buddhist Temples. *Chin. J. Publ. Health* 19: 138–143.
- Knize, M.G., Salmon, C.P., Pais, P. and Felton, J.S. (1999). Food Heating and the Formation of Heterocyclic Aromatic Amine and Polycyclic Aromatic Hydrocarbon Mutagens/ Carcinogens. *Adv. Exp. Med. Biol.* 459: 179–193.
- Laupeze, B., Amiot, L., Sparfel, L., Ferrec, E.L., Fauchet, R. and Fardel, O. (2002). Polycyclic Aromatic Hydrocarbons affect Functional Differentiation and Maturation of Human Monocyte-derived Dendritic Cells. *J. Immunol.* 168: 2652–2658.
- Li, C.S. and Ro, Y.S. (2000). Indoor Characteristics of Polycyclic Aromatic Hydrocarbons in the Urban Atmosphere of Taipei. *Atmos. Environ.* 34: 611–620.
- Lin, Y.M., Zhou, S.Q., Shih, S.I., Lin, S.L., Wang, L.C. and Wu, Z.S. (2011). Fate of Polychlorinated Dibenzop-dioxins and Dibenzofurans during the Thermal Treatment of Electric Arc Furnace Fly Ash. *Aerosol Air Qual. Res.* 11: 584–595.
- Lo, Y.Y., Wang, I.C., Lee, M.L. and Chou, M.S. (2011). Removal of Particulates from Emissions of Joss Paper Furnaces. *Aerosol Air Qual. Res.* 11: 429–436.
- Löfroth, G., Stensman, C. and Brandhorst-Satzkorn, M. (1991). Indoor Sources of Mutagenic Aerosol Particulate Matter: Smoking, Cooking and Incense Burning. *Mutat. Res.* 261: 21–28.
- Lu, H., Zhu, L. and Chen, S. (2008). Pollution Level, Phase Distribution and Health Risk of Polycyclic Aromatic Hydrocarbons in Indoor Air at Public Places of Hangzhou, China. *Environ. Pollut.* 152: 569–575.
- Lung, S.C., Mao, I.F. and Liu, L.J. (2007). Residents' Particle Exposures in Six Different Communities in Taiwan. *Sci. Total Environ.* 377: 81–92.
- Mannix, R.C., Nguyen, K.P., Tan, E.W., Ho, E.E. and Phalen, R.F. (1996). Physical Characterization of Incense Aerosols. *Sci. Total Environ.* 193: 149–158.
- McAughy, J.J. (1997). *Regional Lung Deposition And Dose of ambient particulate in Humans by Particle Mass and Number*, Research Report, AEA Technology, Aerosol Sci. Centre, Oxfordshire (UK).
- Navasumrit, P., Arayasiri, M., Hiang, O.M., Leechawengwongs, M., Promvijit, J., Choonvisase, S., Chantchaemsai, S., Nakngam, N., Mahidol, C. and Ruchirawat, M. (2008). Potential Health Effects of Exposure to Carcinogenic Compounds in Incense Smoke in Temple Workers. *Chem. Biol. Interact.* 173: 19–31.
- Ning, Z. and Sioutas, C. (2010). Atmospheric Processes Influencing Aerosols Generated by Combustion and the Inference of Their Impact on Public Exposure: A Review. *Aerosol Air Qual. Res.* 10: 43–58.
- Nisbet, C. and LaGcy, P. (1992). Toxic Equivalency Factors (TEFs) for Polycyclic Aromatic Hydrocarbons (PAHs). *Regul. Toxicol. Pharm.* 16: 290–300.
- Page, T.J., O'Brien, S., Jefcoate, C.R. and Czuprynski, C.J. (2002). 7, 12-Dimethylbenz[*a*]anthracene Induces Apoptosis in Murine Pre-B Cells through a Caspase-8-dependent Pathway. *Mol. Pharmacol.* 62: 313–319.
- Pope, C.A., 3rd, Thun, M.J., Namboodiri, M.M., Dockery, D.W., Evans, J.S., Speizer, F.E. and Heath Jr., C.W. (1995). Particulate Air Pollution as a Predictor of Mortality in a Prospective Study of U.S. Adults. *Am. J. Respir. Crit. Care Med.* 151: 669–674.
- U.S. Environmental Protection Agency (USEPA). (2002). Health Assessment Document for Diesel Engine Exhaust, EPA/600/8–90/057F, Office of Research and Development.
- Van Grevenynghe, J., Rion, S., Le Ferrec, E., Le Vee, M., Amiot, L., Fauchet, R. and Fardel, Olivier. (2003). Polycyclic Aromatic Hydrocarbons Inhibit Differentiation of Human Monocytes into Macrophages. *J. Immunol.* 170: 2374–2381.
- Voutsas, D. and Samara, C. (2002). Labile and Bioaccessible Fractions of Heavy Metals in the Airborne Particulate



- Matter from Urban and Industrial Areas. *Atmos. Environ.* 36: 3583–3590.
- Wu, Y.L., Li, H.W., Chien, C.H., Lai, Y.C. and Wang, L.C. (2010). Monitoring and Identification of Polychlorinated Dibenzo-*p*-dioxins and Dibenzofurans in the Ambient Central Taiwan. *Aerosol Air Qual. Res.* 10: 463–471.
- Yang, C.R., Lin, T.C. and Chang, F.H. (2006). Correlation between Calcium Carbonate Content and Emission Characteristics of Incense. *J. Air Waste Manage. Assoc.* 56: 1726–1732.
- Yang, C.R., Lin, T.C. and Chang, F.H. (2007). Particle Size Distribution and PAH Concentrations of Incense Smoke in a Combustion Chamber. *Environ. Pollut.* 145: 606–615.
- Yang, C.Y., Chiu, J.F., Cheng, M.F. and Lin, M.C. (1997). Effects of Indoor Environmental Factors on Respiratory Health of Children in a Subtropical Climate. *Environ. Res.* 75: 49–55.
- Yousef, S.E., Brendan, J.M., Dixon, D.G. and Bruce, M.G. (2002). Measurement of Short- and Long-term Toxicity of Polycyclic Aromatic Hydrocarbons Using Luminescent Bacteria. *Ecotoxicol. Environ. Saf.* 51: 12–21.

*Received for review, September 14, 2011*

*Accepted, March 15, 2012*